REGULAR ARTICLE

Assimilation dynamics of soil carbon and nitrogen by wheat roots and Collembola

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Received: 12 August 2006/Accepted: 23 April 2007/Published online: 26 May 2007 © Springer Science+Business Media B.V. 2007

Abstract It has been demonstrated that plant roots can take up small amounts of low-molecular weight (LMW) compounds from the surrounding soil. Root uptake of LMW compounds have been investigated by applying isotopically labelled sugars or amino acids but not labelled organic matter. We tested whether wheat roots took up LMW compounds released from dual-labelled (¹³C and ¹⁵N) green manure by analysing for excess ¹³C in roots. To estimate the fraction of green manure C that potentially was available for root uptake, excess ¹³C and ¹⁵N in the primary decomposers was estimated by

Responsible Editor: A. C. Borstlap.

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Department of Freshwater Ecology, National Environmental Research Institute, Vejlsøvej 25, DK-8600 Silkeborg, Denmark analysing soil dwelling Collembola that primarily feed on fungi or microfauna. The experimental setup consisted of soil microcosm with wheat and duallabelled green manure additions. Plant growth, plant N and recoveries of ¹³C and ¹⁵N in soil, roots, shoots and Collembola were measured at 27, 56 and 84 days. We found a small (<1%) but significant uptake of green manure derived ¹³C in roots at the first but not the two last samplings. About 50% of green manure C was not recovered from the soil-plant system at 27 days and additional 8% was not recovered at 84 days. Up to 23% of C in collembolans derived from the green manure at 56 days (the 27 days sampling was lost). Using a linear mixing model we estimated that roots or root effluxes provided the main C source for collembolans (54-79%). We conclude that there is no solid support for claiming that roots assimilated green manure derived C due to very small or no recoveries of excess ¹³C in wheat roots. During the incubation the pool of green manure derived C available for root uptake decreased due to decomposition. However, the isotopic composition in Collembola indicated that there was a considerable fraction of green manure derived C in the decomposer system at 56 days thus supporting the premise that LMW compounds containing C from the green manure was released throughout the incubation.

Keywords Dissolved organic carbon \cdot Green manure \cdot ¹⁵N \cdot ¹³C \cdot Linear mixing model \cdot *Folsomia fimetaria* \cdot *Proisotoma minuta* \cdot *Protaphorura armata*

Introduction

Plant roots have a large capacity to take up of lowmolecular weight (LMW) compounds through the proton-coupled transporter system, whose primary function is to recapture compounds released from the roots (Farrar et al. 2003). In addition to recapturing root exudates, it has been documented that roots can take up LMW compounds such as amino acids (Nasholm et al. 2001; Lipson and Nasholm 2001) and B-vitamins (Mozafar 1994). Root uptake of LMW compounds has primarily been investigated using isotopically labelled glucose, amino acids or CO₂ (Boddy et al. 2007; Johnson et al. 1996; Kuzyakov and Jones 2006; Owen and Jones 2001). In these experiments very little (2-10%) of the labelled C was recovered in roots indicating that the capture of C from decomposing organic matter in soil does not have a significant impact on the plant's C budget (Kuzyakov and Jones 2006).

Most studies have investigated root uptake of LMW compounds with pulse injection (Bardgett et al. 2003; Nasholm et al. 2001; Weigelt et al. 2005) but to our knowledge only very few studies have used ¹³C or ¹⁴C labelled organic matter for the purpose of detecting root uptake of LMW compounds (Jones et al. 2005). Hodge et al. (1999, 2000) used duallabelled organic matter in experiments with sedge grass but neither study found root uptake of C that derived from the added organic matter. However, their experimental setup was designed for different purposes than monitoring root uptake of C. Application of labelled organic matter might hold some advantages over pulse injection for monitoring root uptake of MSW compounds. (1) The compounds are released continuously and gradually over a longer period and (2) in local hot spots where the density of roots might be larger than in the surrounding, less nutrient rich soil. Compounds released in the rhizospehere are more likely to be taken up than those released outside the rhizosphere. In contrast, direct injection might favour microbial capture because the compounds are spread uniformly throughout the rhizosphere with the possibility that the significance of root C uptake is underestimated (Kuzyakov and Jones 2006).

The C cycle is closely linked to the N cycle but mineralization and immobilization processes lead to decoupling between C and N through respiration of CO₂, volatilisation of NH₃-N or trans- and deamination in amino acids (Magid et al. 2004). The decoupling might lead to over- or underestimation of the importance of root C uptake in long-term labelling studies if the sources of metabolically active C cannot be distinguished from each other. In studies using ¹³C and ¹⁵N labelled additions C and N decomposition processes can be estimated through isotopic analysis of soil fauna living on primary decomposers (Coleman et al. 2002; Garrett et al. 2001). Sampling and analysis of soil fauna is relatively simpler than that of micro-organisms and is especially suited for investigating mass fluxes in long-term studies (Albers et al. 2006).

In this study, we used a readily decomposable green manure enriched with ¹³C and ¹⁵N added to soil microcosms with Collembola and wheat (Triticum aestivum). Wheat was chosen due to its documented ability of taking up exogenously applied LMW compounds (Martens and Frankenberger 1994; Nasholm et al. 2001; Owen and Jones 2001). We tested whether organic C from the green manure was taken up by roots as it has been documented with injection of LMW compounds. Because ¹³C labelled compounds are respired during incubation, the available pool of labelled LMW compounds are likely to decrease with time. To estimate the pool size of the metabolically active C derived from the green manure we analysed the isotopic values of three collembolan species: Proisotoma minuta (Tullberg), Protaphorura armata (Tullberg) and Folsomia fimetaria (L.) that were added at the initiation of the experiment. We used collembolans as "sampling devices" because they have very fast tissue turnover (Bakonyi 1998; Garrett et al. 2001) and are closely linked to the rhizosphere channel through their feeding on rhizosphere microbes and microfauna (Ostle et al. 2007). We expected that *P. armata* would have isotopic values closer to roots than the two other species because herbivory has been reported for P. armata and the closely related Protaphorura fimata but not for F. fimetaria or P. minuta (Brown 1985; Joosse and Koelman 1979; Ulber 1983). We further tested whether the presence of Collembola would affect plant growth and uptake of N since previous studies have yielded contradictory results on this matter (e.g. Cole et al. 2004; Filser 2002; Partsch et al. 2006).

Materials and methods

Soil, green manure and Collembola

A low-nutrient coarse sandy soil was collected from Jyndevad research station in early June 2004, Denmark. The soil was sieved through an 8 mm meshsize sieve, freeze-thawed at -20 and +20 °C in three consecutive cycles over 2 weeks to remove the mesoand macrofauna and subsequently pre-incubated for 21 days at 20-25 °C before use. Soil water content was 14% (38% of gravimetric water holding capacity) and pH was 6.8. Additional properties of the soil are found under No Green Manure (NGM) soil in Table 1. The NGM soil is also referred to as the soil resource. The green manures consisted of ryegrass (Lolium perenne L.). The unlabelled ryegrass was grown at the National Environmental Research Institute in Silkeborg, Denmark and is henceforth referred to as NERI ryegrass. The ryegrass seedlings were grown outside between 1 June and 9 July 2004, in peat moss added clay granules and NPK (Stenrøgel Mosebrug A/S). An additional 50 mg urea l^{-1} peat moss was added 20 days after planting. The average temperature was 15.5 $^{\circ}$ C and the day length 17 h. The ¹³C and ¹⁵N enriched ryegrass was kindly provided by Plant Research International, Wageningen The Netherlands, and is henceforth referred to as Wageningen ryegrass. The ryegrass seedlings were grown for 35 days in a sandy loam soil fertilized with 79.6 mg NH₄NO₃ containing 7.6 mg 15 NH₄⁺ in 0.65-L pots in growth chambers, in which a continuously ¹³C-labelled atmosphere was maintained (Gorissen et al. 1996). This setup ensured that all compounds in the labelled plant material were homogenously labelled. Growth conditions were: light 16 h, PAR 300 μ mol m⁻² s⁻¹, temperature shoot compartment 18/16 °C (day/night), root compartment 16/14 °C (day/night), relative humidity 70%/80% (day/night). After harvest both types of ryegrass were treated in a similar manner with air-drying in an oven with air circulation at 40 °C for 3 days. The ryegrass was ground (1 mm mesh size) to enable homogeneously mixing of the labelled and unlabelled material through the soil. The unlabelled green manure (UGM) consisted only of NERI ryegrass and was weighed into 3.30 g portions for each pot. The labelled green manure (LGM) was weighed into 3.425 g portions consisting of 0.275 g Wageningen

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Tabl	e 1 Content of	soil carbon an	nd nitrogen (mear	$1 \pm SE$, w/GM; n	= 12, NGM;	n = 3)						
Day	% C		% N		C:N (atom	(δ^{13} C (‰)			$\delta^{15}N~(\%0)$		
	w/GM	NGM	w/GM	NGM	w/GM	NGM	LGM	UGM	NGM	LGM	NGM	NGM
0	1.84 ± 0.03	1.52 ± 0.02	0.155 ± 0.002	0.111 ± 0.001	13.9 ± 0.0	16.0 ± 0.1	16.0 ± 0.5	-22.6 ± 0.0	-22.7 ± 0.0	220 ± 2	5.0 ± 0.1	6.5 ± 0.5
27	1.71 ± 0.02	1.56 ± 0.00	0.143 ± 0.002	0.113 ± 0.000	13.9 ± 0.0	16.1 ± 0.0	-3.6 ± 0.1	-22.5 ± 0.1	-23.0 ± 0.0	163 ± 1	5.3 ± 0.1	5.4 ± 0.0
56	1.70 ± 0.03	1.43 ± 0.02	0.134 ± 0.002	0.102 ± 0.001	14.8 ± 0.0	16.3 ± 0.0	-4.8 ± 0.4	-22.7 ± 0.1	-23.1 ± 0.1	126 ± 2	5.3 ± 0.1	5.9 ± 0.2
84	1.66 ± 0.02	1.54 ± 0.01	0.128 ± 0.002	0.112 ± 0.001	15.2 ± 0.1	16.1 ± 0.0	-6.8 ± 0.1	-22.8 ± 0.1	-22.6 ± 0.1	116 ± 1	5.1 ± 0.1	6.6 ± 0.2

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	% C	% N	C: N (atom)	δ^{13} C	δ^{15} N
Unlabelled GM	43.5 ± 0.1	6.78 ± 0.02	7.48 ± 0.0	-24.3 ± 0.0	0.07 ± 0.24
Labelled GM	42.8 ± 0.0	6.59 ± 0.04	7.58 ± 0.0	216 ± 2	827 ± 5

Table 2 Carbon, nitrogen and staple isotopes in green manures (*GM*) (mean \pm SE n = 3)

ryegrass and 3.15 g NERI ryegrass. The Wageningen ryegrass was used in small quantities due to the high costs of the material. Isotope, C and N-values of the labelled and UGMs are found in Table 2. The Wageningen ryegrass was analysed indirectly as the enrichments were too high for the analytical range of the mass spectrometer. Wageningen ryegrass characteristics were, on basis of the measurements of labelled and UGM, determined to: C : N-ratio 7.91 \pm 0.05, 3,655 \pm 2 δ^{13} C, 15,389 \pm 4 δ^{15} N (mean \pm SE, n = 3). The collembolan species *P. minuta, P. armata* and *F. fimetaria* were obtained from laboratory cultures that had been raised on yeast for more than ten generations.

Experimental design

The microcosm setup consisted of 42 pots with three green manure treatments and two Collembola treatments: (1) Unlabelled Green Manure with Collembola (UGM with Collebola), (2) Unlabelled Green Manure without Collembola (UGM without Collembola), (3) Labelled Green Manure with Collembola), (4) Labelled Green Manure without Collembola), (4) Labelled Green Manure without Collembola) and (5) No Green Manure without Collembola (NGM). There were three replicates for each of the treatments. All pots were sown with wheat (*T. aestivum* L. cultivar Vinjett). Pots were sampled destructively three times at different plant ages: 27, 56 and 84 days.

The microcosms were contained in opaque PE cylinders with an inner diameter of 7.1 cm and height of 14.4 cm. The bottom and the top of each pot were closed with PVC lids. An acrylic cylinder with an inner diameter of 2.6 cm and height of 5.0 cm was inserted vertically through a hole in the top lid. All interfaces were sealed with gas tight silicone. An inand outlet tube for air (PVC 4.0×0.8 mm) was inserted into the headspace of each cylinder and the outlet tube was connected to a vacuum pump continuously leading headspace air out of the growth

chamber. The pots were placed in a 16 °C climate chamber with 14:10 h light/dark cycle PAR 400 μ mol m⁻² s⁻¹. Each portion of the green manure, 3.425 g LGM and 3.30 g UGM, was mixed into 547 g soil [dry weight (DW) equivalent] with a ladle. From this mixture, 50 g (DW equivalent) were collected for soil analysis. The 500 g soil (DW equivalent) and green manure remaining in each pot was equivalent to 126.3 kg DW m⁻². The amounts of C and N from the green manures were equivalent to 363 g C m^{-2} and 51.4 g N m⁻² in the UGM treatment 370 g C m⁻² and 51.8 g N m⁻² in the LGM treatment. With an estimated N efficiency rate of $\sim 25\%$ (Janzen and Schaalje 1992), the N application rate was similar to field application. After sampling, the soil was compacted to a height of 10 cm $(1.25 \text{ g DW cm}^{-3})$. Then, 20 individuals of each of the three collembolan species were added to the soil surface (Collembola treatments only). Three wheat seeds were sown in each of the acrylic cylinders right after soil addition. After 5 days, the most viable plant was retained (equal to 253 plants m^{-2}) and the rest were removed. After 7 days, a 1 cm layer of a two-component silicone (Silastic 3,481, Dow Corning) was placed around each plant to seal the pots. Pots were watered with distilled water twice a week or more frequently when needed (max 40 ml water loss per pot) by adding water until the initial weight of the pot was reached. The initial weight was gradually adjusted for plant growth. Water addition was monitored throughout the 84 days that elapsed during the incubation in the five treatments in addition to the three pots without plants. Water loss was followed in all treatments to estimate water use efficiency for wheat.

Termination and analyses

Plants were collected at each sampling occasion around 11.00 AM, and divided into shoots and roots. Leaf fall during incubation was collected and pooled with the subsequent plant sampling. Roots were thoroughly washed to ensure that all soil particles were removed. All plant materials were dried immediately at 60 °C until constant weight and then ballmilled (1 mm mesh size). The soil samples (20 g) were dried at 105 °C for 24 h and then ball-milled. About 4 mg plant material and 40 mg soil, respectively, were weighed into tin cups and analysed for the C, N, ¹³C and ¹⁵N with a Europa Scientific 20-20 IRMS at the Plant and Soil Science laboratory in Copenhagen. Mineral N (NO₃⁻ and NH₄⁺) was measured on fresh soil samples in 1 M KCl, which were shaken and filtered, and subsequently analysed on a flow injection system by Steins Laboratorium A/S, Denmark. The replicates sampled for the mineral N analyses were pooled for each treatment to obtain sufficient material. Collembola were extracted from ~ 150 g soil (DW equivalent) with a Macfadyen high-gradient extractor. The animals were extracted into plastic containers with plaster of Paris and activated charcoal. The containers were kept at a constant 4 °C during the 5 days of extraction. After the extraction collembolans were separated into species and transferred to new substrates without activated charcoal and kept for 24 h at 20 °C to empty their guts. The animals were frozen $(-18 \degree C)$ for 24 h and subsequently dried at 45 °C for 6 h. If any sample from a treatment contained <75 µg biomass the samples were pooled to obtain sufficient N for detection. The animals were analysed with a dual inlet IRMS Europa Hydra 20-20 at UC Davis Stable Isotope Facility, Department of Plant Sciences, California. The isotopic values of the Plant and Soil Science Laboratory were adjusted to those from UC Davis using standards of ethanenitrile.

Statistics and calculations

The isotopic ratios are reported with units of per mil (‰) difference according to:

$$\delta_{\text{sample}} = (R_{\text{sample}}/R_{\text{std}} - 1) \times 1000, \tag{1}$$

where δ is ¹³C or ¹⁵N, R_{sample} is the ratio of the rare to abundant isotope, R_{std} refers to the isotope standards, $R_{\text{C}} = 0.011180$ (VPDB) for ¹³C/¹²C and air, $R_{\text{N}} = 0.0036765$ (air) for ¹⁵N/¹⁴N (Fry 2006). Assuming that two phases, the sink and its source, is in equilibrium, the isotopic fractionation α is described as:

$$\alpha = R_{\rm sink}/R_{\rm source} = (1000 + \delta_{\rm sink})/(1000 + \delta_{\rm source})$$
(2)

and the isotopic enrichment is

$$\varepsilon = (\alpha - 1) \times 1000. \tag{3}$$

The fraction f of C or N derived from the labelled element in a sample is:

$$f_{\text{sample}} = \left(\delta_{\text{sample}} - \delta_{\text{background}}\right) / \left(\delta_{\text{labelled}} - \delta_{\text{background}}\right),$$
(4)

where $\delta_{\text{background}}$ is the isotopic value of the system without the labelled material and δ_{labelled} is the isotopic value of the labelled material. The mass of labelled derived C or N in the system, l_{sample} :

$$l_{\text{sample}} = f_{\text{sample}} \times m_{\text{sample}}, \tag{5}$$

where m_{sample} is the mass of total C or N in the system. l_{sample} is used to calculate the total recovery r of labelled C and N in the system according to:

$$r_{\text{sample}}^{t=1} = l_{\text{sample}}^{t=1} / l_{\text{sample}}^{t=0}, \tag{6}$$

where t = 0 is time at initiation and t = 1 is the time of sampling.

A linear mixing model according to Phillips and Gregg (2003) was applied to the isotope values of Collembola in the LGM treatment to distinguish between the contribution of C and N from three sources: roots, soil resource and LGM. This approach was used instead of directly applying mixing models since it provides all the source combinations that may explain a given sink isotopic composition within given tolerance values. The source increment was set to 0.01 and the tolerance of variability in δ^{13} C and δ^{15} N-values to 4‰. The fractionation values used in the model were taken from laboratory reared Collembola on yeast diet (Table 3, day 0). The isotope values of the yeast were $-16.6\% \delta^{13}$ C and -0.1% δ^{15} N. C : N ratios of the three sources were not taken into account because Collembola were assumed not to feed directly on any of these sources but on the primary decomposers.

All statistical analyses were performed with S-Plus Version 6.2. Student's *t*-test and ANOVA (untransformed data, tested for normal distribution) were used

Day	Species	% C	N %	C:N (atom)	δ^{13} C (‰) ^b		$\delta^{15}N$ (%0)		μg individual ⁻¹	${\rm Mg}~{\rm dw}~{\rm m}^{-2}$	Cumulated release ^c
		w/GM	w/GM	w/GM	LGM	lab/UGM	LGM	lab/UGM			mg $\rm NH_4^+-N~m^{-2}$
0	P. minuta	48.4 ± 0.1	11.2 ± 0.1	5.02 ± 0.04		-16.6 ± 0.1		3.6 ± 0.3	4.14 ± 0.22	20.9 ± 1.1	
	P. armata	53.2 ± 0.3	9.81 ± 0.12	6.32 ± 0.06		-17.4 ± 0.1		2.9 ± 0.7	10.0 ± 0.9	50.5 ± 4.4	
	F. fimetaria	49.3 ± 0.1	9.91 ± 0.11	5.80 ± 0.06		-16.1 ± 0.1		3.0 ± 0.0	5.02 ± 0.58	25.4 ± 2.9	
56	P. minuta	48.4 ± 0.2	9.58 ± 0.09	5.90 ± 0.03	34.7	-21.2	574	-1.0	1.34 ± 0.12	41.2 ± 8.0	5.9
	P. armata	49.0 ± 0.8	10.5 ± 0.3	5.44 ± 0.16	-4.6 ± 0.6	-20.7 ± 0.1	599 ± 4	0.0 ± 1.2	4.23 ± 0.50	162 ± 21	6.6
	F. fimetaria	49.0 ± 0.7	8.74 ± 0.31	6.54 ± 0.24	20.3	-21.2	562	0.1	1.77 ± 0.18	55.7 ± 16.3	7.8
84	P. minuta	48.4 ± 0.7	9.61 ± 0.10	5.88 ± 0.09	12.2	-21.0 ± 0.1	565	-2.3 ± 2.1	1.37 ± 0.12	82.1 ± 12.0	21.6
	P. armata	48.0 ± 0.4	11.3 ± 0.2	4.97 ± 0.06	-4.8 ± 1.2	-21.4 ± 0.2	588 ± 15	3.2 ± 0.3	6.26 ± 0.33	524 ± 46	47.3
	F. fimetaria	49.2 ± 0.1	9.04 ± 0.53	6.34 ± 0.28	12.4	-21.6 ± 0.1	521	-0.9 ± 2.5	1.73 ± 0.14	85.5 ± 27.2	20.2
w/G/	A soils with lab	elled and unla	abelled green 1	nanure, LGM lal	belled green m	anure soils, UC	<i>FM</i> unlabelle	ed green manu	rre soils		
^a Va	lues from day () are from lab	oratory anima	ls							

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to detect differences unless otherwise noted. Wilcoxon rank sum tests were applied when the errors were not normally distributed. Precision of the sample means are shown with standard error of means. In the isotope mixing models, errors were propagated according to Fry (2006).

Results

Wheat

Samples without ±SE were pooled from three samples to obtain sufficient biomass for analysis

Estimated

The presence of Collembola did not affect shoot biomass (Collembola × sampling day interaction, $F_{2,12} = 0.04$, P = 0.96), N content ($F_{2,12} = 0.31$, P = 0.74) or δ^{15} N ($F_{2,12} = 1.32$, P = 0.30). Therefore, GM treatments with and without Collembola were pooled. Likewise, shoot biomass (GM treatment × sampling day interaction, $F_{2,30} = 1.33$, P = 0.28) and N content ($F_{2,30} = 2.84$, P = 0.07) were similar between both LGM and UGM treatments and these parameters were therefore pooled (Table 4). Shoots in the NGM treatment had a much lower growth reaching about 6% of the DW of green manure treatments at 84 days.

 δ^{13} C-values of LGM roots were significantly higher than UGM roots at 27 days but only by 0.9‰ (Fig. 1b; Wilcox, P < 0.05). If it is assumed that this was due to uptake of LGM-C, the percentage derived from LGM C (%Cdf LGM) would be 0.38 ± 1.16% (Table 5). This was the only occasion with significant difference between the labelled and unlabelled treatments for roots as well as shoots (Fig. 1a, b). In contrast, δ^{13} C-values of NGM shoots and roots were 2.7–7.1‰ lower than those of UGM. This difference was correlated with the water use efficiency that was around three times lower per unit dry matter in NGM compared to the GM treatments at 27 days and 4.5 times lower at 56 and 84 days.

Plant δ^{15} N-values in the LGM treatment were between 670 and 750‰ indicating that most of the N derived from the added GM. In the shoots, between 87 and 90% of N derived from LGM (%Ndf LGM) and in the roots between 82 and 85% (Table 5). Shoots contained 5.7 ± 0.3% less N from soil resources than roots. The amount of LGM-N taken up by plants was 42% at 84 days. This amount is underestimated because of root loss during washing. The only source of N for plants in the NGM soil was

	Day	% C		% N		C : N (atom)		$g dw m^{-2}$		(00%) N _{C1} 3	
		w/GM	NGM	w/GM	NGM	w/GM	NGM	w/GM	NGM	$\mathrm{UGM}^{\mathrm{a}}$	NGM
Shoots	27	40.8 ± 0.1	37.8 ± 1.1	5.54 ± 0.07	4.10 ± 0.22	8.59 ± 0.08	10.8 ± 0.48	176 ± 5	26.1 ± 5.1	0.4 ± 0.5	0.1 ± 0.3
	56	44.2 ± 0.1	37.8 ± 1.7	1.96 ± 0.04	2.39 ± 0.24	26.3 ± 0.5	18.5 ± 0.9	983 ± 21	66.5 ± 6.6	2.8 ± 0.5	0.2 ± 0.8
	84	44.2 ± 0.0	41.0 ± 0.8	1.37 ± 0.04	1.77 ± 0.30	37.7 ± 0.9	27.0 ± 3.4	$1,792 \pm 47$	101 ± 13	2.2 ± 0.5	-1.1 ± 0.3
Roots	27	41.9 ± 0.5	42.7 ± 0.3	2.57 ± 0.14	2.23 ± 0.27	19.0 ± 0.9	22.4 ± 2.7	47 ± 5	20.2 ± 7.7	-1.5 ± 0.8	-4.5 ± 0.3
	56	44.0 ± 0.2	42.6 ± 0.2	1.49 ± 0.09	1.31 ± 0.06	34.5 ± 2.0	38.0 ± 1.0	149 ± 5	17.7 ± 3.9	-0.6 ± 0.6	-4.2 ± 0.3
	84	44.4 ± 0.2	43.7 ± 0.2	1.15 ± 0.05	1.29 ± 0.07	44.9 ± 1.7	39.4 ± 1.4	118 ± 6	20.2 ± 5.3	-1.5 ± 0.3	-4.7 ± 0.2

Estimated values

Fable 4 Content of wheat carbon, nitrogen, biomass (dry weight) yield and isotopic difference, $s^{15}N$ (‰) relative to sources (mean \pm SE, w/GM n = 12, UGM n = 6, NGM

that of the soil resource. Consequently, the isotopic difference, ε^{15} N, between plant parts and soil resource could be readily determined (Table 4). On average, NGM roots were depleted by 4.2 ± 0.4‰ compared to NGM shoots. In the UGM treatment ε^{15} N-values were estimated on the basis on the fraction of plant N that derived from soil resource N and LGM-N in the LGM treatment. On average, LGM roots were depleted by 2.9 ± 0.6‰ compared to LGM shoots (Table 4). Addition of GM also increased the utilization of the soil resource N compared to NGM except for the first plant sampling at day 27. At 84 days NGM plants had taken up 2.05 g N from soil resources versus 3.50 g N in the LGM treatment.

Collembola

Collembola samples from day 27 were lost due to a breakdown of the extractor. At 56 and 84 days the δ^{15} N-values in the LGM treatment were between 520 and 600‰ and the δ^{13} C-values between -5 and 35‰ (Table 3). Most of Collembola body N, 64-73%, derived from LGM. In contrast, C mainly derived from soil resources or root exudates with 7 and 24% from LGM-C (Table 5). On Fig. 2, the isotopic values of Collembola are placed relative to the three sources of C and N: (1) roots, (2) soil resource and (3) LGM. Collembola are placed relatively closer to roots than the two other sources. The linear mixing model by Phillips and Gregg (2003) predicted that from 77 to 79% of the P. armata body mass derived from roots opposed to 54–67% for *P. minuta* (Table 5). The fraction of C and N derived from roots increased between 56 and 84 days. P. armata obtained less of their mass from LGM compared to P. minuta and F. fimetaria. The isotope values of Collembola in the NGM treatment were very close and could not be used to infer sources (Table 3).

The numbers of collembolans were $101,600 \pm 18,100 \text{ m}^{-2}$ at day 56 and $202,900 \pm 13,800 \text{ m}^{-2}$ at day 84. The only significant difference between species was at 84 days where *P. armata* was more abundant than *F. fimetaria* (*P* < 0.05). Excretion of NH₄⁺-N was estimated from collembolan biomass (Table 3) and the ammonium excretion values by Larsen et al. (2007). The estimated accumulated excretion of NH₄⁺-N at 84 days was 89 mg N m⁻². This amount corresponded to





about 0.3% compared to total plant N uptake at 84 days (Table 4).

Recoveries of LGM-C and -N

Recovery of LGM-C from soil was 49% at 27 days. At 84 days, recovery of LGM-C had had decreased to 41% (P < 0.01, Fig. 3). Total recovery of LGM-N included soil and plant N and was similar at 27,

56 and 84 days with 90% (P < 0.05, Fig. 3). The following values of mineral N were found at day 0: For NO₃⁻-N, 3.16, 2.78 and 1.52 g m⁻² for the LGM, UGM and NGM treated soils, respectively; for NH₄⁺-N, 6.69, 6.44 and 0.417 g m⁻² the LGM, UGM and NGM treated soils, respectively. Thus, an amount of ~10 g N m⁻² was readily available in mineral form and predominantly NH₄⁺-N right after addition to soil, which was about one-fifth of total GM added.

Table 5 Percentages of C and N derived from labelled green manure (*df* LGM) and estimated food sources from linear mixing modelling (mean \pm SE, plant; n = 6, Collembola; n = 3)

	Day	Elements		Linear mixi	ng model	
		%C df LGM	%N df LGM	Roots	Soil resource	LGM
Shoots	27	0	89.7 ± 0.0			
	56	0	90.4 ± 0.0			
	84	0	87.2 ± 0.0			
Roots	27	0.38 ± 1.16	85.0 ± 0.0			
	56	0	85.5 ± 0.0			
	84	0	81.5 ± 0.0			
P. minuta	56	23.4 ± 0.2	69.9 ± 0.3	54 ± 1	23 ± 0	23 ± 1
	84	13.9 ± 0.1	69.0 ± 0.3	67 ± 2	19 ± 0	14 ± 1
P. armata	56	6.7 ± 0.5	72.9 ± 0.0	77 ± 2	16 ± 0	7 ± 1
	84	6.9 ± 0.9	71.2 ± 0.2	79 ± 2	14 ± 1	7 ± 1
F. fimetaria	56	17.4 ± 0.1	68.4 ± 0.3	59 ± 2	24 ± 1	17 ± 1
	84	14.2 ± 0.1	63.5 ± 0.3	60 ± 1	26 ± 0	14 ± 1



Fig. 2 Source mixing triangle for Collembola. The numbers next to symbols signify sampling day

Discussion

Fig. 3 Total recovery of

(LGM) C (a) and N (b) from plant plus soil (mean \pm SE,

labelled green manure

n = 6

The study was designed to test whether wheat roots would take up ¹³C derived from the dual-labelled green manure, LGM-C. LGM roots were enriched more than UGM roots at 27 days but the isotopic difference was very small (P < 0.05, $\sim 1\%$) and we estimated the amount of C derived from LGM to be 0.38 ± 1.16% (Table 5). We cannot account for the fraction of LGM-C taken up by roots that was subsequently respired but Kuzyakov and Jones (2006) found this to be 45% for maize roots that had taken up exogenously applied glucose. Thus, the amount of LGM-C taken up by roots would still

represent <1% of total plant C. There is a possibility that the small root uptake of exogenously C was an artefact because of contamination of the root surface with labelled C. However, the ¹³C signatures in roots at the two last samplings were similar and it supports the premise that the washing procedure neatly cleaned off labelled particles from the root surface. The identical ¹³C signatures of the LGM and UGM shoots show that the difference in roots cannot be attributed to different water use efficiencies as for the NGM plants (Farquhar and Richards 1984). Because of the very small or total lack of recovery of excess 13 C in wheat roots, we do not find that there is solid support for claiming that roots assimilated green manure derived carbon. This supports findings from injection studies that uptake of exogenously applied LMW compounds is small or insignificant indicating that roots are poor competitors for LMW compounds (Boddy et al. 2007; Despland and Noseworthy 2006; Kuzyakov and Jones 2006; Owen and Jones 2001).

Most of plant N derived from the green manure but apparently there was a smaller fraction of LGM-N in roots than in shoots (Table 5). This variation can partly be explained by differences in isotope values between the two plant compartments observed at natural abundance. In the UGM treatment, roots were estimated to be depleted by about 3‰ compared to shoots, which also agrees with previous natural abundance studies (Evans et al. 1996; Lopes and Araus 2006; Lopes et al. 2004). Assuming a similar



fractionation in the LGM treatment, only about oneeight of the depletion of ¹⁵N in LGM roots can be explained by intra-plant fractionation. Another explanation for the isotopic differences between the two plant compartments could be that roots assimilated N at a later stage than shoots when the soil was more depleted in ¹⁵N. However, the similar ¹⁵N-values at 27-56 days (Fig. 1c, d) do not support the notion of differential assimilation events between the two plant compartments. So the intra-plant variation was probably caused by different patterns of assimilation. For example, significant intra-plant variation has been observed when NO_3^- is the primary nitrogen source, but little variation has been observed when NH₄⁺ is the source (Evans et al. 1996; Yoneyama et al. 1991). NH_4^+ is assimilated immediately in the root why organic nitrogen in shoots and roots is a product of a single assimilation event. In contrast, NO₃⁻ assimilation that can occur in both roots and shoots (Evans 2001).

Collembola were primarily included as sampling devices of primary decomposers. The isotopic values of Collembola from 56 and 84 days depict that their tissue N mainly derived from LGM and that C mainly derived from roots or soil resources (Fig. 2). For P. minuta that do not eat plant roots (Ulber 1983), the fraction of LGM-C decreased from 23 to 14% between 56 and 84 days showing that C and N from LGM was increasingly but not completely decoupled in the decomposer system. The linear mixing model revealed that C from roots, the rhizosphere channel, was very important for Collembola in general and P. armata in particular (Table 5). This agrees with our hypothesis that P. armata would derive more of it resources from roots than the two other species because it can feed directly on roots. Our finding that the rhizosphere channel was the main supplier of energy in the soil food web is in line with previous findings (Garrett et al. 2001; Pelz et al. 2005; Ruf et al. 2006). However, the large loss of LGM-C at the beginning of the experiment indicates that LGM-C was very important for fuelling the microbial mediated release of mineral N and subsequent plant growth. Collembola generally incorporated more C derived from soil resources than from LGM. In contrast, Collembola showed a clear preference for newly incorporated C from litter over the soil resource in the study by Briones et al. (1999) without plants and with a slower decomposable green manure.

In spite of comparable or higher collembolan densities than at field conditions (Axelsen and Thorup-Kristensen 2000), neither plant biomass nor N content were not affected by the presence of Collembola. This is in line with Cole et al. (2004) who found no effect by Collembola on plant productivity in monocultures. Collembolan release of mineral N was estimated to 0.3% compared to plant N uptake and it is evident that the direct contribution by Collembola was insignificant. Collembolan effects are often ascribed to their grazing on fungal hyphae (Lussenhop 1996) or spreading of fungal propagules (Filser 2002) but fungi might have been less important for decomposition than bacteria in this study because the organic matter was readily decomposable through a rapid release and utilization of soluble carbon compounds (Van Ginkel and Gorissen 1998). Moreover, the fertilization level was high compared to normal agricultural settings because the 42% uptake the added LGM-N was higher than we expected. Thus, the ample supply of mineral N reduced the importance of collembolan mediated processes.

Magid et al. (2004) proposed that the decomposition of intracellular LMW compounds and proteins can be viewed as a process separate from the decomposition of macro-polymers in cell walls. In our study, these two processes seemed to be separated with an initial rapid loss of C that followed by a slower loss of C. N was also lost within the first 27 days. The loss was most likely due to a rapid breakdown of the proteins in the green manure or ammonium volatilisation due to senescence of microorganisms (Hogberg 1997). It is unlikely that denitrification occurred since the soil was not water logged. The coincidence of C and N losses early in this study indicate that microbial decomposition decoupled substantial amounts of labelled compounds from the LGM giving the micro-organisms a pre-emptive competitive advantage in the acquisition of easily available LGM resources.

We stated in the introduction that application of labelled organic matter might hold an advantage over pulse injection of labelled compounds because LMW compounds are released gradually during the decomposition processes. It can be argued, however, that there was no gradual release of LMW compounds in this study as a lot of C was lost in the initial phase before the onset of plant growth. However, according to the linear mixing model more than one-fifth of P. minuta body mass derived from LGM at 56 days indicating that not all labelled compounds from LGM were decoupled within the first 27 days. Also, an additional 8% of LGM-C was lost between 27 and 84 days showing that the remaining pool of C was recycled and released continuously. The sensitivity of detecting uptake of LGM-C in roots would have been higher if the enrichment of ¹³C in LGM had been higher than in this study. But the conclusion that root uptake of C from exogenously applied organic matter was very small or insignificant would probably remain the same. However, the methodology of using ¹³C and ¹⁵N LGM in a plant-soil system in combination with isotopic analysis of soil fauna provided a lot of valuable information on C and N dynamics in soil and the significance of plant derived C for the soil food web.

Acknowledgements This study was part of the research project "Closing the Rural-Urban Nutrient Cycle (CRUCIAL)" supported by Danish Research Centre for Organic Farming, DARCOF. Marc Ventura was supported by a Marie Curie post-doctoral grant (MEIF-CT-2005-010554). We thank Ole Andersen from Sejet Planteforædling for supplying the wheat seeds.

References

- Albers D, Schaefer M, Scheu S (2006) Incorporation of plant carbon into the soil animal food web of an arable system. Ecology 87(1):235–245
- Axelsen JA, Thorup-Kristensen K (2000) Collembola and mites in plots fertilised with different types of green manure. Pedobiologia 44:556–566
- Bakonyi G (1998) Nitrogen turnover of *Sinella coeca* (Collembola: Entomobryidae). Eur J Entomol 95(3):321–326
- Bardgett RD, Streeter TC, Bol R (2003) Soil microbes compete effectively with plants for organic-nitrogen inputs to temperate grasslands. Ecology 84(5):1277–1287
- Boddy E, Hill P W, Farrar J, Jones DL (2007) Fast turnover of low molecular weight components of the dissolved organic carbon pool of temperate grassland field soils. Soil Biol Biochem 39(4):827–835
- Briones MJI, Ineson P, Sleep D (1999) Use of δ^{13} C to determine food selection in collembolan species. Soil Biol Biochem 31(6):937–940
- Brown RA (1985) Effects of some root-grazing arthropods on the growth of sugar-beet. In: Fitter AH, Atkinson D, Read DJ, Usher MB (eds) Ecological interactions in soil: plants, microbes and animals. Blackwell Science Published, Oxford, pp 285–295
- Cole L, Staddon PL, Sleep D, Bardgett RD (2004) Soil animals influence microbial abundance, but not plant-microbial

competition for soil organic nitrogen. Funct Ecol 18(5):631–640

- Coleman D, Fu S, Hendrix P, Crossley D Jr (2002) Soil foodwebs in agroecosystems: impacts of herbivory and tillage management. Eur J Soil Biol 38(1):21–28
- Despland E, Noseworthy M (2006) How well do specialist feeders regulate nutrient intake? Evidence from a gregarious tree-feeding caterpillar. J Exp Biol 209(7):1301– 1309
- Evans RD (2001) Physiological mechanisms influencing plant nitrogen isotope composition. Trends Plant Sci 6(3):121–126
- Evans RD, Bloom AJ, Sukrapanna SS, Ehleringer JR (1996) Nitrogen isotope composition of tomato (*Lycopersicon esculentum* Mill. cv. T-5) grown under ammonium or nitrate nutrition. Plant Cell Environ 19(11):1317–1323
- Farquhar GD, Richards RA (1984) Isotopic composition of plant carbon correlates with water-use efficiency of wheat genotypes. Aus J Plant Physiol 11(6):539–552
- Farrar J, Hawes M, Jones D, Lindow S (2003) How roots control the flux of carbon to the rhizosphere. Ecology 84(4):827–837
- Filser J (2002) The role of Collembola in carbon and nitrogen cycling in soil. Pedobiologia 46(3–4):234–245
- Fry B (2006) Stable Isotope Ecology. Springer, Berlin, p 308
- Garrett CJ, Crossley DA Jr, Coleman DC, Hendrix PF, Kisselle KW, Potter RL (2001) Impact of the rhizosphere on soil microarthropods in agroecosystems on the Georgia piedmont. Appl Soil Ecol 16(2):141–148
- Gorissen A, Kuikman PJ, Van Ginkel JH, Van de Beek H, Jansen AG (1996) ESPAS—an advanced phytotron for measuring carbon dynamics in a whole plant-soil system. Plant Soil 179(1):81–87
- Hodge A, Robinson D, Griffiths BS, Fitter AH (1999) Why plants bother: root proliferation results in increased nitrogen capture from an organic patch when two grasses compete. Plant Cell Environ 22(7):811–820
- Hodge A, Stewart J, Robinson D, Griffiths BS, Fitter AH (2000) Spatial and physical heterogeneity of N supply from soil does not influence N capture by two grass species. Funct Ecol 14(5):645–653
- Hogberg P (1997) Tansley review No 95—N-15 natural abundance in soil-plant systems. New Phytol 137(2):179–203
- Janzen HH, Schaalje GB (1992) Barley response to nitrogen and non-nutritional benefits of legume green manure. Plant Soil 142(1):19–30
- Johnson JF, Allan DL, Vance CP, Weiblen G (1996) Root carbon dioxide fixation by phosphorus-deficient Lupinus albus—contribution to organic acid exudation by proteoid roots. Plant Physiol 112(1):19–30
- Jones DL, Healey JR, Willett VB, Farrar JF, Hodge A (2005) Dissolved organic nitrogen uptake by plants—an important N uptake pathway? Soil Biol Biochem 37(3):413–423
- Joosse ENG, Koelman T (1979) Evidence for the presence of aggregation pheromones in *Onychiurus armatus* (Collembola), a pest insect in sugar-beet. Entomol Exp Appl 26(2):197–201
- Kuzyakov Y, Jones DL (2006) Glucose uptake by maize roots and its transformation in the rhizosphere. Soil Biol Biochem 38(5):851–860
- Larsen T, Luxhøi J, Magid J, Jensen LS, Krogh PH (2007) Properties of anaerobically digested and composted mu-

nicipal solid waste assessed by linking soil mesofauna dynamics and nitrogen modelling. Biol Fertil Soils DOI 10.1007/s00374-007-0178-x

- Lipson D, Nasholm T (2001) The unexpected versatility of plants: organic nitrogen use and availability in terrestrial ecosystems. Oecologia 128(3):305–316
- Lopes MS, Araus JL (2006) Nitrogen source and water regime effects on durum wheat photosynthesis and stable carbon and nitrogen isotope composition. Physiol Plantarum 126(3):435–445
- Lopes MS, Nogues S, Araus JL (2004) Nitrogen source and water regime effects on barley photosynthesis and isotope signature. Funct Plant Biol 31(10):995–1003
- Lussenhop J (1996) Collembola as mediators of microbial symbiont effects upon soybean. Soil Biol Biochem 28(3):363–369
- Magid J, Luxhøi J, Lyshede OB (2004) Decomposition of plant residues at low temperatures separates turnover of nitrogen and energy rich tissue components in time. Plant Soil 258(1):351–365
- Martens DA, Frankenberger WT (1994) Assimilation of exogenous 2'-14C-indole-3-acetic acid and 3'-14C-tryptophan exposed to the roots of three wheat varieties. Plant Soil 166(2):281–290
- Mozafar A (1994) Enrichment of some B-vitamins in plants with application of organic fertilizers. Plant Soil 167:305–311
- Nasholm T, Huss-Danell K, Hogberg P (2001) Uptake of glycine by field grown wheat. New Phytol 150(1):59–63
- Ostle N, Briones MJI, Ineson P, Cole L, Staddon P, Sleep D (2007) Isotopic detection of recent photosynthate carbon flow into grassland rhizosphere fauna. Soil Biol Biochem 39(3):768–777
- Owen AG, Jones DL (2001) Competition for amino acids between wheat roots and rhizosphere microorganisms and

the role of amino acids in plant N acquisition. Soil Biol Biochem 33(4–5):651–657

- Partsch S, Milcu A, Scheu S (2006) Decomposers (Lumbricidae, Collembola) affect plant performance in model grasslands of different diversity. Ecology 87(10):2548– 2558
- Pelz O, Abraham WR, Saurer M, Siegwolf R, Zeyer J (2005) Microbial assimilation of plant-derived carbon in soil traced by isotope analysis. Biol Fertil Soils 41(3):153–162
- Phillips DL, Gregg JW (2003) Source partitioning using stable isotopes: coping with too many sources. Oecologia 136(2):261–269
- Ruf A, Kuzyakov Y, Lopatovskaya O (2006) Carbon fluxes in soil food webs of increasing complexity revealed by ¹⁴C labelling and ¹³C natural abundance. Soil Biol Biochem 38(8):2390–2400
- Ulber B (1983) Einfluss von Onychiurus fimatus Gisin (Collembola, Onychiuridae) and Folsomia fimetaria L. (Collembola, Isotomidae) auf Pythium ultimum Trow., einen erreger des Wurzelbrandes der Zuckerrübe. In: Lebrun P, André HM, De Medts A, Grégoire Wibo C, Wauthy G (eds) New trends in soil biology. Proceeding of VIII. International Coll. Soil Zool. Louvain-la-Neuve (Belgium), pp 261–268, August 30–September 2, 1982
- Van Ginkel JH, Gorissen A (1998) In situ decomposition of grass roots as affected by elevated atmospheric carbon dioxide. Soil Sci Soc Am J 62(4):951–958
- Weigelt A, Bol R, Bardgett RD (2005) Preferential uptake of soil nitrogen forms by grassland plant species. Oecologia 142(4):627–635
- Yoneyama T, Omata T, Nakata S, Yazaki J (1991) Fractionation of nitrogen isotopes during the uptake and assimilation of ammonia by plants. Plant Cell Physiol 32(8):1211–1217