# 1 Cellulose fermentation by the gut microbiota is likely

# 2 not essential for the nutrition of millipedes

#### 3 Running title: Gut microbiota in millipede nutrition

- 4 Julius Eyiuche Nweze<sup>a,b</sup>, Shruti Gupta<sup>a\*</sup>, Michaela M. Salcher<sup>a</sup>, Vladimír Šustr<sup>a</sup>, Terézia
- 5 Horváthová<sup>c\*</sup>, Roey Angel<sup>a,b #</sup>
- 6 <sup>a</sup>Institute of Soil Biology and Biogeochemistry, Biology Centre CAS, České Budějovice,
- 7 Czechia
- 8 <sup>b</sup>Faculty of Science, University of South Bohemia in České Budějovice, Czechia
- 9 °Institute of Hydrobiology, Biology Centre CAS, České Budějovice, Czechia

10

- 11 Email addresses: julius.nweze@bc.cas.cz, microbioshruti86@yahoo.in,
- 12 <u>michaelasalcher@gmail.com</u>, <u>vladimir.sustr@bc.cas.cz</u>, <u>tereza.horvathova@gmail.com</u>,
- 13 roey.angel@bc.cas.cz
- 14 \* Present addresses:
- 15 Shruti Gupta Faculty of Science, University of South Bohemia in České Budějovice,
- 16 Czechia
- 17 Terézia Horváthová Department of Aquatic Ecology, EAWAG, Dübendorf, Switzerland
- 18 # Corresponding author: roey.angel@bc.cas.cz
- 19 Words: abstract: 215 words; main text: 4534 words
- 20
- 21

#### 22 Abstract

Millipedes are thought to depend on their gut microbiome for processing plant-litter-23 24 cellulose through fermentation, similar to many other arthropods. However, this hypothesis lacks sufficient evidence. To investigate this, we disrupted the gut microbiota of juvenile 25 Epibolus pulchripes (tropical, CH<sub>4</sub>-emitting) and Glomeris connexa (temperate, non-CH<sub>4</sub>-26 emitting) using chemical inhibitors and isotopic labelling. Feeding the millipedes sterile or 27 antibiotics-treated litter notably reduced faecal production and microbial load without major 28 29 impacts on survival or weight. Bacterial diversity remained similar, with Bacteroidota 30 dominant in *E. pulchripes* and *Pseudomonadota* in G. connexa. Sodium-2-31 bromoethanesulfonate treatment halted CH4 emissions and reduced the faecal mcrA 32 copies in *E. pulchripes* after 14 days, but emissions resumed after returning to normal 33 feeding. Methanogens in the order Methanobacteriales and Methanomasscilliicoccales 34 associated with protists were detected using Catalysed Reporter Deposition Fluorescence In situ Hybridization (CARD-FISH) on day 21, despite suppressed CH<sub>4</sub>-emission. 35 36 Employing <sup>13</sup>C-labeled leaf litter and RNA-SIP revealed a slow and gradual prokaryote 37 labelling, indicating a significant density shift only by day 21. In addition to labelling of taxa 38 from orders well-recognized for their role in (ligno)cellulose fermentation (e.g., 39 Bacteroidales, Burkholderiales, and Enterobacterales), others, such as members of Desulfovibrionales were also labelled. Surprisingly, labelling of the fungal biomass was 40 somewhat guicker. Our findings suggest that fermentation by the gut microbiota is likely 41 not essential for the millipede's nutrition. 42

#### 43 Importance

44 Millipedes (Diplopoda) constitute the third most significant group of detritivores after 45 termites and earthworms, yet they have been comparatively understudied. Traditionally, it was believed that millipedes gain energy from fermenting cellulose using their gut 46 47 microbiota, similar to wood-feeding termites, but this belief lacks evidence. This study used 48 two model millipede species, *Epibolus pulchripes* (large, tropical, and methane emitter) 49 and Glomeris connexa (small, temperate, and non-methane emitter) to test this belief. We 50 used chemical manipulation experiments, stable isotope labelling, and DNA sequencing to comprehend the microbiota's role in the millipede's nutrition. The findings suggest that 51 52 cellulose fermentation by the gut microbiota may not be essential for millipede nutrition; 53 instead, bacteriovory and fungivory might be the dominant feeding strategies of millipedes.

# 54 Introduction

Like most animals, invertebrates form intricate partnerships with diverse microbial communities (1), contributing significantly to their evolutionary and ecological success (2). This close interconnectedness has led to the concept of animals as "holobionts," where the host and its microbiota are viewed as a single ecological entity (3, 4). Recent studies on microbiomes provide further evidence of the widespread prevalence of microbial partnerships across the animal kingdom (5, 6).

While most invertebrates have microbial associations, their reliance on them varies widely. 61 Termites, for instance, depend entirely on their gut microbiota for nutrition (7). Conversely, 62 63 many other arthropods, such as caterpillars, may lack a resident gut microbiota and develop fully even germ-free (8). Most arthropods generally fall between these extremes, 64 65 relying on their microbiota for some form of support (e.g. cockroaches (9, 10) or isopods (11, 12)). Detritivorous and xylophagous animals typically rely on gut microorganisms for 66 67 cellulose digestion. Although animal cellulases are found in some gut systems (13). (ligno)cellulolytic bacteria, fungi and protists are generally deemed necessary for 68 69 hydrolysis and fermentation, releasing short-chain fatty acids, which get absorbed by the 70 host (14).

Millipedes (Diplopoda) are crucial detritivores widely distributed and abundant in many 71 72 temperate and tropical ecosystems (15). Despite their status as keystone species in 73 tropical and temperate forests (16), millipedes have been understudied compared to other detritivores, particularly concerning their microbiome. Due to the nutrient-poor nature of 74 plant litter, millipedes compensate for low assimilation efficiencies through high ingestion 75 rates (17). Similar to other arthropods, millipedes host diverse gut microorganisms (18). 76 77 Notably, the central hindgut was shown to host the highest microorganism density, attaching to its cuticle, while the foregut and midgut contain mostly transient inhabitants 78 79 (19). Various studies suggest that certain millipede gut bacteria possess enzymes for breaking down plant polysaccharides (20-24). If millipedes rely on cellulose for their 80 nutrition, extensive fermentation followed by methanogenesis, similar to ruminants or 81 wood-feeding termites, should occur in their guts (7, 25). However, methanogenesis has 82 only been observed in some millipede species, but not others, with its occurrence 83 84 correlated to the millipede size (26). Despite these findings, direct proof of gut 85 microorganisms supporting the millipede's nutritional needs has not been experimentally

86 demonstrated. An alternative hypothesis suggests millipedes foster microbial growth in 87 litter, potentially digesting the resulting fungal and bacterial biomass (27).

To investigate the role of the millipede gut microbiota, we conducted experiments using two model species: the CH<sub>4</sub>-emitting *Epibolus pulchripes* (Spirobolida) and *Glomeris connexa* (Glomerida), which do not emit CH<sub>4</sub>. *E. pulchripes* is a large millipede (130–160 mm) common along the East African coast (28), while *G. connexa* is smaller (10-17 mm) and native to Central Europe (29). We assessed the impact of inhibitors on body weight, survival, faecal bacterial load, gut bacterial composition, and CH<sub>4</sub> production. Additionally, we identified metabolically active hindgut prokaryotes using <sup>13</sup>C-RNA-SIP.

95

# 96 **Results**

## 97 Effect of antibiotic curing

Feeding millipedes with either sterile or antibiotics-treated feed led to only negligible 98 99 weight change in both species (Fig 1a and b; Table S2) with no significant trend. The 100 treatment also did not significantly impact the millipedes' survival based on Kaplan-Meier estimates (Fig. S1). Despite maintaining a stable weight, faecal production decreased over 101 102 time in response to antibiotics or sterile feed (P < 2.2e-16 for both species; Fig. 1c and d; Table S3). No significant difference was found between the treated groups in E. 103 pulchripes, but in G. connexa the sterile-litter group was different from the antibiotic-104 treated groups (*P* < 0.0001). Total faecal colony counts in both millipede species were also 105 106 consistently higher in the control group compared to the antibiotic-treated or sterile feeding groups at all time points (P < 0.0001; Fig. 1e and f; Table S4). After 35 days for E. 107 108 pulchripes and 16 days for G. connexa, most animals in the treatment groups ceased faecal production, leading to the cessation of plate counts. Once again, only the sterile-109 110 litter group in G. connexa differed from the other treatment groups. Total faecal 16S rRNA gene copies in *E. pulchripes* were reduced by 61%–77% in the treated groups compared 111 112 to the control group (P = 0.01), while In G. connexa, 34%–74% reductions were observed in the treated groups (P < 0.001; Fig. 1g; Table S5). In both species, no difference 113 between the treated groups was observed. After noting a substantial decrease in bacterial 114 115 load, we measured CH<sub>4</sub> emission on day 35 (Fig. 1h; Table S6). As anticipated, CH<sub>4</sub> was present in *E. pulchripes* but absent in *G. connexa* (data not shown). The control groups 116 displayed a significantly higher CH<sub>4</sub> production rate (284.1 ± 58 nmol mg<sup>-1</sup> d<sup>-1</sup>) than the 117 other treatments (P = 0.0008). However, the treated groups saw a 57–74% reduction in 118 CH<sub>4</sub> production without significant differences between them. 119

## 120 Prokaryotic community compositions after treatment

We sequenced 48 samples of *E. pulchripes* and *G. connexa*, consisting of 12 hindguts and 12 faecal samples for each species. The average sequencing depth stood at ca. 40K reads per sample, post-processing of reads and decontamination (Table S7 and S8). The two millipede species differed remarkably in their microbial composition, with the phylum *Bacteroidota* dominating the hindgut of *E. pulchripes* and *Pseudomonadota* that of *G. connexa*. In each case, these phyla comprised over 50% of the abundance regardless of

127 treatment (Fig. 2a and b; Table S9). Pseudomonadota, Bacteroidota and Actinobacteriota

128 dominated both species' faecal pellets. (Fig. 2c and d). On the genus level, *E. pulchripes*'

129 hindgut and faecal samples were primarily dominated by Citrobacter, Bacteroides, and

130 Pseudomonas (Fig. 2e-h; Table S9). In contrast, G. connexa showed differences between

131 hindgut and faecal sample compositions, with faecal samples appearing more diverse (Fig.

132 **2h)**.

## **Impact of treatment on prokaryotic community structures**

134 Overall, no significant differences were found in alpha diversity within or between

135 treatment groups in the hindguts (Fig. 3a & b; Fig. S2; Table S10) or faeces (Fig. 3c & d;

136 Fig. S2; Table S10) of *E. pulchripes* and *G. connexa*. *E. pulchripes*' hindgut groups

137 displayed greater bacterial diversity and richness than *G. connexa*. Constrained analysis

138 of principal coordinates (CAP) revealed significant differences in microbial community

139 composition among sterile feeding or antibiotics treatments in both hindguts and faeces of

140 both species (Fig. 3e, f, g & h; Table S10). ANCOM-BC2 analysis identified only a handful

141 of microbial genera with differential abundance between treatments (Table S11; Fig. S3),

142 indicating that the antibiotic treatment worked relatively non-selective. The few taxa with a

143 decrease in the mean absolute abundance (e.g. *Streptomycetaceae* and *Mucilaginibacter* 

144 from the *E. pulchripes*' faeces) are known to often possess antibiotic resistance genes.

## 145 Influence of BES inhibition on methanogenesis in *E. pulchripes*

146 Na-BES-treated litter was provided to investigate the importance of methanogenesis in the

147 CH<sub>4</sub>-emitting *E. pulchripes*. Methane emissions showed no significant differences on days

148 0 (P = 0.19) and 7 (P = 0.08; Fig. 4A; Table S12). However, by day 14, CH<sub>4</sub> production

149 was nearly fully inhibited (P =  $2.7 \times 10^{-4}$ ) and remained so for an additional 21 days. Upon

switching to untreated litter on day 35, CH<sub>4</sub> emissions began recovering on day 49 and

151 resumed pre-treatment values by day 63. Despite some average weight increase in the

152 treated groups, no significant difference was detected at any time (Fig. 4b).

After inhibiting methane production for 21 days, a suspension made from fresh faeces was examined under a bright-field microscope, revealing various protists, nematodes, and rotifers ranging from 12 to 100  $\mu$ m in size (Fig. S4). The ciliate abundance averaged 3 × 10<sup>5</sup> ml<sup>-1</sup>, regardless of treatment (Fig. 4c; Table S13). Quantification of the *mcrA* gene, pivotal in methanogenesis (30), showed a significant reduction in the two Na-BES-treated groups compared to the control (P = 0.02; Fig. 4d; Table S13). CARD-FISH was used to

- detect the presence of free-living (Fig. S5) and symbiotic archaea (Fig. S6), primarily
- 160 methanogens, in protists from faecal samples. The amplicon sequencing data indicated
- 161 that members of the *Methanomassciillicoccales* and *Methanobacteriales* were the
- 162 dominant methanogens in *E. pulchripes*, and these orders were accordingly targeted.
- 163 Although *mcrA* copy numbers declined, positive hybridisation signals for these
- 164 methanogen orders were observed in both Na-BES treatments. Methanogens were
- 165 detected on the 0.2  $\mu m$  filter (Fig. S5) and associated with protists as endosymbionts (Fig.
- 166 4e; Fig. S6), with no significant changes in its count per ciliate (Fig. 4f).

## 167 Detection of active microbiota with <sup>13</sup>C-RNA-SIP

- 168 RNA-SIP was used to identify the active microorganisms in the millipedes' gut on a
- 169 temporal scale (Table S14). The shift in peak of 16S rRNA towards the denser gradient
- 170 fractions, indicating label incorporation, was evident by day 3 and more prominently by day
- 171 7 for *E. pulchripes* and day 14 for *G. connexa* (Fig. 5). Nevertheless, despite feeding on
- 172 fully-labelled litter for 21 days, a significant portion of RNA remained unlabelled.
- 173 Surprisingly, the labelling of the fungal biomass, represented by the 18S rRNA peak,
- 174 shifted earlier towards denser gradient fractions compared to 16S rRNA in both millipede
- species (Fig. S7). However, the lack of pronounced peak deviation compared to the
- 176 control in some replicates and days does not necessarily imply unsuccessful labelling
- 177 since the labelled fraction of the community might still be too small. Indeed, there was a
- significant change in community composition in the heavy fractions of labelled gradients
- 179 compared to unlabelled ones already by day 3 (Fig. S8; Table S15).
- 180 For comparing heavy fractions in labelled versus unlabelled gradients of 16S RNA, an
- 181 average of  $1305 \pm 59$  and  $579 \pm 41$  ASVs were used for *E. pulchripes* and *G. connexa* per
- time point after filtering (Table S16). Surprisingly, the model identified, on average, only
- 183 22% of the ASVs in *E. pulchripes* and 24% in *G. connexa* as labelled. These values were
- 184 consistent over time. Therefore, the shift in copy-number peaks towards denser fractions,
- as observed in Fig. 5, was due to increased labelling in already labelled ASVs rather than
- 186 a change in the proportion of labelled ASVs.

## 187 Diversity of active microbiota in a heavy fraction of <sup>13</sup>C-RNA-

#### 188 **SIP**

- 189 In agreement with the general bacterial diversity in the gut, the major phyla whose
- 190 members were flagged as labelled were Actinobacteriota, Bacillota, Bacteroidota, and
- 191 Pseudomonadota (Fig. 6; Table S16). In E. pulchripes, Bacillota comprised 35 to 55.3%,
- 192 Bacteroidota 13.1 to 15.1% and Pseudomonadota from 13.8 to 23% of the total labelled
- ASVs. In *G. connexa*, *Bacillota* comprised 20.4 to 45.9% of total significant ASVs,
- 194 Pseudomonadota ranged from 20 to 51.6%, Actinobacteriota from 15.1% to 22.6%, and
- 195 Bacteroidota from 3.2 to 10.8%. Fig. S9-15 show the phylogenetic distribution of the
- 196 labelled ASVs across the samples in each of the major bacterial classes. Despite our
- 197 expectation for gradual labelling of the microorganisms with time, similar numbers and, in
- 198 many cases, the same ASVs were consistently labelled throughout the incubation. In *E.*
- 199 pulchripes, members of the classes Clostridia and the orders Bacteroidales, Rhizobiales,
- 200 Enterobacterales, Desulfovibrionales, Pirellulales, Verrucomicrobiales and Victivallales
- 201 were most prominently labelled. In G. connexa, members of the class Clostridia and the
- 202 orders Bacteroidales, Rhodobacterales, Enterobacterales, Pseudomondales and
- 203 Micrococcales were most prominently labelled.

204

# 205 **Discussion**

206 The gut microbiota, crucial for the ecophysiology of arthropods (31), is especially vital for

- 207 detritivores relying on recalcitrant plant polymers with low nitrogen content. Building on
- culture-based (32) and recent molecular studies (21, 23, 33, 34), the findings underscore a
- 209 generally stable and species-specific millipede gut microbiota, resistant to inhibitors.
- 210 Variations in closely related arthropods may arise from gut conditions like pH, oxygen
- availability (35), and gut topography (19). Specifically for millipedes, hindgut volume
- 212 differences, influencing redox potential, likely contribute to microbiota variations, promoting
- fermentation and methanogenesis in larger species (e.g., *E. pulchripes* and *T. aoutii*) but
- not in smaller ones (e.g., *G. connexa*) (21, 26).
- 215 Curing or sterilizing arthropods to assess their dependence on gut microbiota has been
- 216 conducted in various species, yielding diverse outcomes. Not surprisingly, for wood-
- 217 feeding termites, exposure to high oxygen levels results in the disappearance of
- flagellates, leading to starvation (7, 36). This is because wood-feeding termites rely on
- short-chain fatty acids, which are the products of cellulose fermentation, for their nutrition.
- 220 Cured arthropods in other studies exhibited moderate responses, including decreased
- feeding and altered microbiota, observed in desert millipedes (24), Carabidae members
- 222 (37), and egg-hatching cockroaches (38). In contrast, larval Lepidoptera, exclusively
- 223 feeding on fresh leaves and likely relying on simple sugars, showed no physiological
- response to antibiotic curing (39). Both millipede species in this study maintained a stable
- weight, suggesting they might not require fermentation products for nutrition. However, the
- 226 notable decrease in faecal production and the relatively unchanged taxonomic composition
- indicated a potentially significant role in the microbiota. Notably, there was a shift in
- 228 abundance towards antibiotic-resistant bacterial strains, such as Citrobacter and
- 229 Bacteroides in E. pulchripes (40, 41) and Pseudomonas and Achromobacter in G.
- 230 connexa (42, 43).
- 231 This study validated CH<sub>4</sub> release in *E. pulchripes*, aligning with previous findings (26, 44).
- 232 Antibiotics decreased CH<sub>4</sub> emission, likely disrupting bacterial fermentation, a
- phenomenon observed in cockroaches when bacteria and flagellates were targeted (45).
- As expected, the application of BES, a specific methanogenesis inhibitor (46), reduced
- 235 CH<sub>4</sub> production to undetectable levels without apparent effects on *E. pulchripes* fitness. As
- 236 CH<sub>4</sub> production serves as a hydrogen sink in anaerobic systems driving syntrophic
- 237 fermentation processes (47), it supports the notion that gut fermentation is non-essential

238 for millipede nutrition. The dominant methanogens, *Methanobacteriales* and

239 Methanomassiliicoccales, in our millipedes are known gut inhabitants (44). Surprisingly,

240 despite suppressed CH4 production and a 10-fold drop in mcrA gene copy numbers,

241 methanogen density in the gut remained unaffected. In dynamic gut systems, members

must continue to proliferate to avoid being flushed out, methanogens likely live as

symbionts of protists, directly benefiting from fermentation products, similar to the case in

244 termites (48, 49).

In the SIP experiment, RNA labelling was slow and gradual, leaving a substantial portion

unlabelled even after 21 days, indicating the inefficiency of the millipede gut system in

247 degrading leaf litter and assimilating carbon. In contrast, fungal biomass exhibited faster

and higher labelling, especially in G. connexa. Soil litter decomposition studies suggest

249 fungi thrive first on recalcitrant and nutrient-poor litter, with bacteria flourishing later on

nutrient-rich and more labile litter (50, 51). In the hindgut of both millipede species (21) and

251 *Telodeinopus aoutii* (23), *Ascomycota* and *Basidiomycota* dominate, mirroring soil

decomposition patterns (50, 52, 53).

253 Despite millipedes' ability to hydrolyze polysaccharides, lipids, and proteins through

salivary gland enzymes alongside their resident microbes (as in many other detritivores;

13, 54, 55) and conditions, and methanogenesis in the digestive tract (26, 44, 56),

cellulose digestion significance in millipede metabolism remains inconclusive. Quantitative

data, including low metabolic rates in millipedes fed pure cellulose, suggest challenges inmaintaining a positive energy balance (57).

259 The labelled microbiota in *E. pulchripes* and *G. connexa*, primarily *Bacillota*, *Bacteroidota*,

and *Pseudomonadota*, show distinctive patterns associated with polysaccharide

degradation, consistent with recent millipede studies (21, 23). Similar labelling of these

262 phyla was observed in a scarab beetle study using 13C-cellulose (58). Although certain

263 labelled taxa (e.g., *Bacteroidales*, *Burkholderiales*, and *Enterobacterales*) are recognized

for their role in (ligno)cellulose fermentation in millipedes (21, 23, 34), others (e.g.,

265 members of Desulfovibrionales) are hindgut microorganisms involved in processes like

sulfate reduction and are likely unrelated to fermentation. Despite senescent leaves not

being exclusively comprised of (ligno)cellulose, these polymers constitute approximately

268 50–75% of litter material (59). In the near absence of other terminal electron acceptors in

the gut, most other simpler carbon sources will also need fermentation for metabolism.

270 Consequently, we conclude that while cellulolytic fermentation occurs in the millipede gut,

it likely contributes minimally to the host's diet.

If fermentation products are not a primary nutritional source for the millipede, their main 272 nutrient origin remains a guestion. Classical <sup>14</sup>C-labelling studies indicated bacterial 273 274 assimilation into the millipede's biomass surpassing that of plants but focused on labgrown strains and omitted fungi (27). Woodlice, another detritivore, exhibit a preference for 275 276 fungi- or bacteria-colonized leaf tissues over natural litter (60, 61). Genomic and transcriptomic screening of the studied millipede species revealed glycoside hydrolases 277 (GH) capable of degrading chitin and peptidoglycan as abundant as, or even more so 278 than, cellulose-degrading GHs (21). The decrease in ergosterol levels post-digestion 279 supports significant fungal digestion in the millipede gut (62). Some species exhibit a 280 281 preference for fungal fruiting bodies, algae, and lichen films (63). Millipedes' midgut fluid 282 effectively kills bacteria in a species-specific manner (64). Coprophagy in millipedes may provide access to fresh microbial and fungal biomass resulting from a partial breakdown of 283 284 recalcitrant plant material (65). Additionally, millipedes produce endogenous GHs in their 285 salivary glands and midgut for digesting non-structural plant material (23, 34, 55). Fluid feeding, described in Colobognatha millipedes, enables feeding on fresh plant material 286 (66). These findings don't exclude other roles of the millipede gut microbiota, such as 287 detoxification of plant toxins (67), essential compound production (23), protection against 288 289 pathogens (33), and even acquiring new genes through horizontal transfer (68). 290 This work demonstrates that cellulose fermentation likely plays a minor role, at best, in the millipede's nutrition. Further work is needed to decipher their exact trophic function in 291 292 nature and the potential role their microbiota plays in their survival and modulating 293 greenhouse gas emissions.

# 294 Materials and Methods

## 295 Animal collection and maintenance

We used juvenile E. pulchripes from our lab breeding colony and wild-caught G. connexa 296 from Czechia (forest near Helfenburk u Bavorova; 49°8'10.32" N, 14°0'24.21" E). No 297 specific permit was required for the collection. Species identification relied on 298 morphological features (69, 70); data not shown). Before use, the animals were kept in the 299 300 lab for several weeks. Both species were housed in perforated plastic terraria, filled with commercial sand as a substrate, broken terracotta pots for shelter, and locally collected or 301 302 purchased Canadian poplar (Populus x canadensis) leaf litter (see below). Moisture (50-60%) was maintained by spraying with tap water every other day. Both species 303 304 experienced a 12-hour photoperiod. E. pulchripes were housed individually in a box (19.3 x 13.8 x 5 cm) at 25 °C and in a climate-controlled room. Meanwhile, five G. connexa 305 306 individuals were kept in each box (15 x 10 x 4 cm) in an incubator (TERMOBOX LBT 165,

307 Vanellus s.r.o.) at a temperature of 15 °C.

#### 308 Antibiotic curing

- 309 Each millipede species comprised 40 individuals split into four groups of ten: Control,
- 310 Sterile, diluted antibiotics (2X-Diluted in *E. pulchripes* and 5X-Diluted in *G. connexa*) and
- 311 undiluted antibiotics (Undiluted in *E. pulchripes* and 2X-Diluted in *G. connexa*). Briefly, the
- 312 Control group was fed untreated, senesced leaves, the Sterile group was fed autoclaved
- 313 leaves, and the antibiotics-treated groups were fed autoclaved leaves treated with
- antibiotics. *E. pulchripes* groups were fed around 2.4 g of litter, while *G. connexa* groups
- received 0.5 g. Just before feeding, the leaf litter was sprayed with 500 µl of tap water
- 316 (Control), sterile distilled water (Sterile), or antibiotics solution containing penicillin G:
- 317 10,000 units ml<sup>-1</sup>, streptomycin sulfate: 10  $\mu$ g ml<sup>-1</sup> and amphotericin B: 25  $\mu$ g ml<sup>-1</sup> (Thermo
- 318 Fisher Scientific), following Zimmer and Bartholme (71). The terraria, sand, and litter were
- 319 replaced weekly to maintain hygiene.
- 320 The animal fitness was followed for 42 days by aseptically measuring their weights (±0.01
- 321 g). During feeding, three fresh faeces pellets (0.15–0.19 g for *E. pulchripes* and 0.01–0.02
- 322 g for *G. connexa*) were sampled from the terraria, suspended in phosphate buffer (2 ml;
- 323 pH 7.4), plated in triplicates on LB-agar plates and incubated at 25 °C. After 16 h, the
- 324 colonies were counted and used to quantify the bacterial load. The remaining faecal

325 material was kept at -20 °C for DNA extraction (see below). Methane emission was also

#### 326 monitored (see below).

## 327 Inhibition of methanogenesis

Thirty E. pulchripes individuals were divided into three groups of ten. The Control group 328 329 was fed on untreated litter, while the other two groups were fed litter treated with 5 mM or 330 10 mM of Sodium 2-bromoethanesulfonate (Na-BES: Sigma-Aldrich) to inhibit 331 methanogenesis. Moisture was maintained by spraying with sterile tap water or Na-BES 332 solution every other day. The animals' weight and CH<sub>4</sub> production were regularly 333 monitored for 64 days. Methane emission measurements were conducted by placing the millipedes in sealed glass bottles with wet filter paper pieces to maintain humidity (130 ml 334 335 bottle for E. pulchripes; 30 ml for G. connexa; Thermo Fisher Scientific) for 4 h at 20 °C. 336 The control was glass vessels without animals. Headspace samples (0.5 ml) were collected at the start and the end of incubation using a gas-tight syringe and analysed on a 337 gas chromatograph (HP 5890 series II; Hewlett Packard) equipped with a 2 m Porapak N 338 column at 75 °C and an FID detector. The difference in CH<sub>4</sub> concentration between start 339 and finish was used to calculate the production rate. 340

#### 341 Identification and enumeration of protists and symbiotic

#### 342 methanogens

Fourteen days post-CH<sub>4</sub>-inhibition, fresh *E. pulchripes* faecal pellets were crushed using a sterilised mortar and pestle, vortexed in 5 ml of 1X phosphate buffer saline (PBS) solution

345 (pH 7.2), and then incubated at room temperature for 2–6 h to dissolve the aggregates.

- 346 After spin-down, 2  $\mu$ l of the supernatant was examined under a bright-field microscope
- 347 (20x) using a Neubauer chamber (Sigma-Aldrich). Protists were identified and
- 348 enumerated. Part of the supernatant was also used for enumerating the ciliate-associated
- 349 archaea and methanogens of the Methanobacteriales and Methanomascilliicoccales
- 350 orders using Catalysed Reporter Deposition Fluorescence in situ Hybridization (CARD-
- 351 FISH; see Supplementary material for further details).

## 352 Stable isotope labelling of RNA

353 For the SIP experiment, three replicates from separate terraria were used for each

- 354 species. *E. pulchripes* had one individual per replicate, while *G. connexa* had five to adjust
- 355 for size differences. Millipedes were fed 99.9% <sup>13</sup>C-labelled Canadian-poplar leaves

356 (IsoLife, Netherlands), Control groups were fed unlabelled leaves, Rearing conditions were maintained as described above. Faecal samples were collected every 2 days for isotopic 357 358 labelling analysis. Then, 1.9 g of faeces from each millipede species were vacuum dried in a SpeedVac DNA130 (Thermo Fisher Scientific) at 45 °C for 3 h, and 25 µg was 359 360 transferred into triplicate tin capsules. Isotopic labelling was assessed at the Stable Isotope Facility, Biology Centre CAS, using a Thermo Scientific<sup>™</sup> 253 Plus<sup>™</sup> 10 kV IRMS 361 equipped with a SmartEA Isolink and GasBench II (Thermo Fisher Scientific). The <sup>13</sup>C at% 362 363 was calculated following Hayes (72: data not shown). Animals were sacrificed and dissected on days 3, 7, 14, and 21 following Sardar et al. (23) and stored at -20 °C for 364 365 subsequent analysis. RNA was extracted from frozen hindgut samples, purified and guantified according to Angel et al. (73). Hindgut samples from the SIP experiment 366 measured 0.677–1.108 g for E. pulchripes and 0.083–0.092 g for G. connexa. See 367

368 Supplementary material for further details.

## 369 Isopycnic ultra-centrifugation of <sup>13</sup>C labelled RNA

- 370 Following RNA purification, density gradient centrifugation was performed in caesium
- trifluoroacetate (CsTFA) density gradients following a previously published protocol (74).
- 372 See Supplementary material for further details.

#### 373 Gene quantification, amplicon library construction and

#### 374 sequencing

375 Pooled faecal pellet samples from the antibiotics curing and inhibition of methanogenesis experiments used for DNA extraction were 0.43-0.59 g for E. pulchripes and 0.20-0.40 for 376 G. connexa. See Supplementary material for further details. DNA extracts from the 377 antibiotics treatment experiment (24 samples per species) were subjected to 16S-rRNA-378 gene guantification using the OX200 AutoDG Droplet Digital PCR System (ddPCR; Bio-379 Rad), primers 338F—805R and the 516P FAM/BHQ1 probe (75). DNA extracts from the 380 381 methanogenesis inhibition experiment were used for quantifying the mcrA gene as a marker for methanogens using primers mlas mod and mcrA-rev, according to Angel et al. 382 383 (76). Before sequencing, the cDNA from the SIP fractions (160 samples for each millipede species) was used for quantifying the 16S rRNA copies of bacteria using the same method 384 385 as mentioned above and the 18S rRNA copies of fungi using the FungiQuant system (77). For amplicon sequencing, the V4 region of the 16S rRNA gene was amplified and 386 sequenced in a two-step protocol on an Illumina MiniSeq platform (Mid Output Kit; 387

Illumina) according to Naqib *et al.* (78). PCR amplification was performed on 10 ng of DNA
or 2 µl of cDNA with primers 515F\_mod and 806R (79), synthesised with the Fluidigm
linkers CS1 and CS2 on their 5' end. Sequencing was performed at the DNA Services
Facility at the University of Illinois, Chicago, USA.

#### 392 **Bioinformatic and statistical analyses**

393 Unless mentioned otherwise, all bioinformatic and statistical analyses were done in R V4.1.1 (80). A linear mixed-effects model (81) was fitted to determine the effect of 394 395 treatments and time on the millipede weight and microbial load. Differences between 396 treatments in terms of total faecal pellet production, methane emission, mcrA and 16S rRNA copies were evaluated using an ANOVA model (82) followed by Tukey's HSD test 397 for pairwise comparisons (83). Survival analysis of the animals was computed using the 398 399 Kaplan-Meier estimates (84). Sequencing data was analysed as follows: primer and linker regions were removed from the raw amplicon reads using Cutadapt (V3.5; 85). The raw 400 401 reads were processed, assembled and filtered using the R package DADA2 (V1.28) with 402 the following non-standard filtering parameters: maxEE = c(2, 2) in the filterAndTrim 403 function and pseudo pooling in the dada function (86). Chimaeras were removed with the removeBimeraDenovo option. The guality-filtered pair-end reads were classified to the 404 405 genus level using SILVA V138 (87), and those not classified as bacteria or archaea were filtered out. Heuristic decontamination was done using the decontam R package (88), and 406 407 unique sequences were identified and clustered in an amplicon sequence variant (ASV) table. The resulting tables were imported into the R package Phyloseg (89). Read counts 408 409 were normalised using median sequencing depth before plotting taxa abundance and after excluding ASVs without taxonomic assignments at the phylum level and those below a 5% 410 prevalence threshold. Alpha diversity indices were computed using the vegan package on 411 unfiltered and non-normalised data (90) and evaluated using the Kruskal-Wallis test (91) 412 and Dunn's test (92). Corrections for multiple testing were made using the Benjamini-413 414 Hochberg (BH; 93) method. Values were compared and converted to a compact letter 415 using the cldList function in the rcompanion package (94). Beta diversity was calculated 416 with a constrained analysis of principal coordinates (CAP; 95). Lastly, a permutational multivariate ANOVA (96); function vegan::adonis) was conducted using the Bray-Curtis 417 distance matrix and the pairwise.adonis2 function (97) to assess combined treatment and 418 pairwise effects on the microbial community. 419

420 Differentially abundant genera were identified after sterile feeding or antibiotic treatment 421 using ANCOM-BC2 (98) after removing all ASVs not present in at least two samples or 422 with an abundance of less than 2. Only genera with adjusted P-values  $\leq$  0.05 and those passing the pseudo-count-addition sensitivity analysis were plotted. 423 424 Identification of isotopically labelled ASVs in the SIP experiment using differential 425 abundance analysis followed Angel (99). After initial processing as described above, rare taxa (with <100 total reads, present in <2 SIP fractions in a given gradient and its 426 427 unlabelled counterpart). The DADA2 output sequences were aligned using sina 1.7.2 (100) against the SILVA V138 DB, and a maximum-likelihood phylogenetic tree was constructed 428 429 using IO-TREE V2.1.1 (101) with the '-fast' option. The 16S rRNA copies were plotted against the density and used to calculate absolute ASV abundances. Fractions with 430 densities >1.795 g ml<sup>-1</sup> ('heavy' fractions) from each labelled sample at each time point 431 were compared against their unlabelled counterparts using DESeg2 V1.40.1 (102), using 432 the parametric fit type and the Wald significance test. Log<sub>2</sub> fold change (LFC) shrinkage 433 was applied using the function lfcShrink (103), and the results were filtered to include only 434 435 ASVs with a positive  $\log_2$  fold change and a p-value <0.1 (one-sided test).

436

## 437 Acknowledgements

We are grateful for the support of Lucie Faktorová and Eva Petrová in collecting G. 438 439 connexa samples, Lucie Faktorová in maintaining the *E. pulchripes* colony and assisting 440 with millipede dissection, and Eva Petrová for her guidance and assistance in DNA and RNA extractions and quantification. We are thankful to Radka Malá for her assistance in 441 442 the filtration and fixation of CARD-FISH samples. Special thanks to Travis Blake Meador, 443 Stanislav Jabinski, and Poláková Ljubov for their contributions to stable isotope detection and guantification in millipede faeces. RA, SG and JEN were supported by a Junior Grant 444 from the Czech Science Foundation (GA ČR), grant number 19-24309Y. The funders had 445 446 no role in study design, data collection and interpretation, or the decision to submit the work for publication 447

# 448 Author Contributions

The approach for this study was conceptualised by RA and VS, experiments were carried out by SG, JEN, MMS and TH, and the data analysis was designed by RA and JEN. The

- 451 bioinformatics analyses were carried out by JEN and RA. The manuscript was written by
- 452 JEN, SG and RA, with significant contributions from MMS and VS. All authors have
- 453 thoroughly reviewed and approved the final version of the manuscript.

## 454 Availability of data and analysis scripts

- 455 The short-read amplicon sequencing data have been deposited under the NCBI BioProject
- 456 PRJNA948469 with BioSample SUB13838396 for antibiotics treatment and SUB13843680
- 457 for RNA-SIP. For reproducibility, reusability, and transparency, the scripts used in this
- 458 study are available on GitHub (https://github.com/ISBB-anaerobic/Active-microbial-
- 459 <u>community-pre-and-post-inhibition.git</u>).

#### 460 **References**

- 1. Petersen JM, Osvatic J. 2018. Microbiomes in nature: Importance of invertebrates in understanding the natural variety of animal-microbe interactions. mSystems 3:e00179-17.
- 2. Moran NA, Ochman H, Hammer TJ. 2019. Evolutionary and ecological consequences of gut microbial communities. Annu Rev Ecol Evol Syst 50:451–475.
- 3. Bordenstein SR, Theis KR. 2015. Host biology in light of the microbiome: ten principles of holobionts and hologenomes. PLoS Biol 13:e1002226.
- 4. Zilber-Rosenberg I, Rosenberg E. 2008. Role of microorganisms in the evolution of animals and plants: the hologenome theory of evolution. FEMS Microbiol Rev 32:723–735.
- Russell JA, Dubilier N, Rudgers JA. 2014. Nature's microbiome: introduction. Mol Ecol 23:1225–1237.
- 6. Vavre F, Kremer N. 2014. Microbial impacts on insect evolutionary diversification: from patterns to mechanisms. Curr Opin Insect Sci 4:29–34.
- Brune A. 2014. Symbiotic digestion of lignocellulose in termite guts. 3. Nat Rev Microbiol 12:168–180.
- Hammer TJ, Sanders JG, Fierer N. 2019. Not all animals need a microbiome. FEMS Microbiology Letters 366:fnz117.
- Mikaelyan A, Thompson CL, Hofer MJ, Brune A. 2016. Deterministic assembly of complex bacterial communities in guts of germ-free cockroaches. Appl Environ Microbiol 82:1256– 1263.
- Tinker KA, Ottesen EA. 2016. The core gut microbiome of the American cockroach, *Periplaneta americana*, is stable and resilient to dietary shifts. Appl Environ Microbiol 82:6603–6610.
- 11. Bouchon D, Zimmer M, Dittmer J. 2016. The terrestrial isopod microbiome: An all-in-one toolbox for animal–microbe interactions of ecological relevance. Front Microbiol 7:1472.
- 12. Mattila JM, Zimmer M, Vesakoski O, Jormalainen V. 2014. Habitat-specific gut microbiota of the marine herbivore *Idotea balthica* (Isopoda). J Exp Mar Biol Ecol 455:22–28.
- 13. Watanabe H, Tokuda G. 2001. Animal cellulases. Cell Mol Life Sci 58:1167–1178.
- 14. Schmidt K, Engel P. 2021. Mechanisms underlying gut microbiota–host interactions in insects. J Exp Biol 224:jeb207696.
- Kime RD, Golovatch SI. 2000. Trends in the ecological strategies and evolution of millipedes (Diplopoda). Biological Journal of the Linnean Society 69:333–349.
- 16. Crawford CS. 1992. Millipedes as model detritivores. Ber nat-med Verein Innsbruck 12.

- 17. David J-F. 2014. The role of litter-feeding macroarthropods in decomposition processes: A reappraisal of common views. Soil Biol Biochem 76:109–118.
- Byzov BA. 2006. Intestinal microbiota of millipedes, p. 89–114. *In* König, H, Varma, A (eds.), Intestinal microorganisms of termites and other invertebrates. Springer-Verlag, Berlin/Heidelberg.
- 19. Nardi JB, Bee CM, Taylor SJ. 2016. Compartmentalization of microbial communities that inhabit the hindguts of millipedes. Arthropod Struct Dev 45:462–474.
- 20. Koubová A, Lorenc F, Horváthová T, Chroňáková A, Šustr V. 2023. Millipede gut-derived microbes as a potential source of cellulolytic enzymes. World J Microbiol Biotechnol 39:169.
- 21. Nweze JE, Šustr V, Brune A, Angel R. 2024. Functional similarity, despite taxonomical divergence in the millipede gut microbiota, points to a common trophic strategy. Microbiome 12:16.
- 22. Ramanathan B, Alagesan P. 2012. Isolation, characterization and role of gut bacteria of three different millipede species. Indian J Sci Res 3:55–61.
- 23. Sardar P, Šustr V, Chroňáková A, Lorenc F, Faktorová L. 2022. *De novo* metatranscriptomic exploration of gene function in the millipede holobiont. 1. Sci Rep 12:16173.
- 24. Taylor EC. 1982. Role of aerobic microbial populations in cellulose digestion by desert millipedes. Appl Environ Microbiol 11.
- 25. Khairunisa BH, Heryakusuma C, Ike K, Mukhopadhyay B, Susanti D. 2023. Evolving understanding of rumen methanogen ecophysiology. Front Microbiol 14.
- 26. Šustr V, Chroňáková A, Semanová S, Tajovský K, Šimek M. 2014. Methane production and methanogenic archaea in the digestive tracts of millipedes (Diplopoda). PLoS one 9.
- Bignell D. 1989. Relative assimilations of <sup>14</sup>C-labelled microbial tissues and <sup>14</sup>C-plant fibre ingested with leaf litter by the millipede *Glomeris marginata* under experimental conditions. Soil Biol Biochem 21:819–827.
- Enghoff H. 2011. East African giant millipedes of the tribe *Pachybolini* (Diplopoda, Spirobolida, Pachybolidae). 1. Zootaxa 2753:1–41.
- 29. Hoess R, Scholl A. 2001. Allozyme and Literature Study of *Glomeris guttata* Risso, 1826, and *G. connexa* Koch, 1847, a Case of Taxonomic Confusion (Diplopoda: Glomeridae). Zool Anz 240:15–33.
- Hedderich R, Whitman WB. 2006. Physiology and biochemistry of the methane-producing Archaea, p. 1050–1079. *In* In: Rosenberg, E., DeLong, E.F., Lory, S., Stackebrandt, E., Thompson, F. (eds) The Prokaryotes. Springer, Berlin, Heidelberg.

- 31. Schapheer C, Pellens R, Scherson R. 2021. Arthropod-Microbiota Integration: Its Importance for Ecosystem Conservation. Frontiers in Microbiology 12.
- Dhivya A, Alagesan P. 2017. Millipedes as Host for Microbes A Review. Int J Microbiol Res 8:19–24.
- Nweze JE, Schweichhart JS, Angel R. 2024. Viral communities in millipede guts: Insights into the diversity and potential role in modulating the microbiome. Environ Microbiol 26:e16586.
- Sardar P, Šustr V, Chroňáková A, Lorenc F. 2022. Metatranscriptomic holobiont analysis of carbohydrate-active enzymes in the millipede *Telodeinopus aoutii* (Diplopoda, Spirostreptida). Front Ecol Evol 10.
- Engel P, Moran NA. 2013. The gut microbiota of insects diversity in structure and function. FEMS Microbiol Rev 37:699–735.
- 36. Ebert A, Brune A. 1997. Hydrogen Concentration Profiles at the Oxic-Anoxic Interface: a Microsensor Study of the Hindgut of the Wood-Feeding Lower Termite Reticulitermes flavipes (Kollar). Appl Environ Microbiol 63:4039–4046.
- 37. Lundgren JG, Lehman RM. 2010. Bacterial gut symbionts contribute to seed digestion in an omnivorous beetle. PLoS One 5:e10831.
- Tegtmeier D, Thompson CL, Schauer C, Brune A. 2016. Oxygen Affects Gut Bacterial Colonization and Metabolic Activities in a Gnotobiotic Cockroach Model. Appl Environ Microbiol 82:1080–1089.
- 39. Hammer TJ, Janzen DH, Hallwachs W, Jaffe SP, Fierer N. 2017. Caterpillars lack a resident gut microbiome. Proc Natl Acad Sci 114:9641–9646.
- 40. Jabeen I, Islam S, Hassan AKMI, Tasnim Z, Shuvo SR. 2023. A brief insight into Citrobacter species a growing threat to public health. Frontiers in Antibiotics 2.
- 41. Rasmussen BA, Bush K, Tally FP. 1993. Antimicrobial Resistance in Bacteroides. Clin Infect Dis 16:S390–S400.
- Abbott IJ, Peleg AY. 2015. *Stenotrophomonas, Achromobacter*, and nonmelioid *Burkholderia* species: antimicrobial resistance and therapeutic strategies, p. 099–110. *In* Seminars in respiratory and critical care medicine. Thieme Medical Publishers.
- Pang Z, Raudonis R, Glick BR, Lin T-J, Cheng Z. 2019. Antibiotic resistance in *Pseudomonas aeruginosa*: mechanisms and alternative therapeutic strategies. Biotechnol Adv 37:177–192.
- 44. Horváthová T, Šustr V, Chroňáková A, Semanová S, Lang K, Dietrich C, Hubáček T, Ardestani MM, Lara AC, Brune A, Šimek M. 2021. Methanogenesis in the Digestive Tracts

of the Tropical Millipedes *Archispirostreptus gigas* (Diplopoda, Spirostreptidae) and *Epibolus pulchripes* (Diplopoda, Pachybolidae). Appl Environ Microbiol 87:e00614-21.

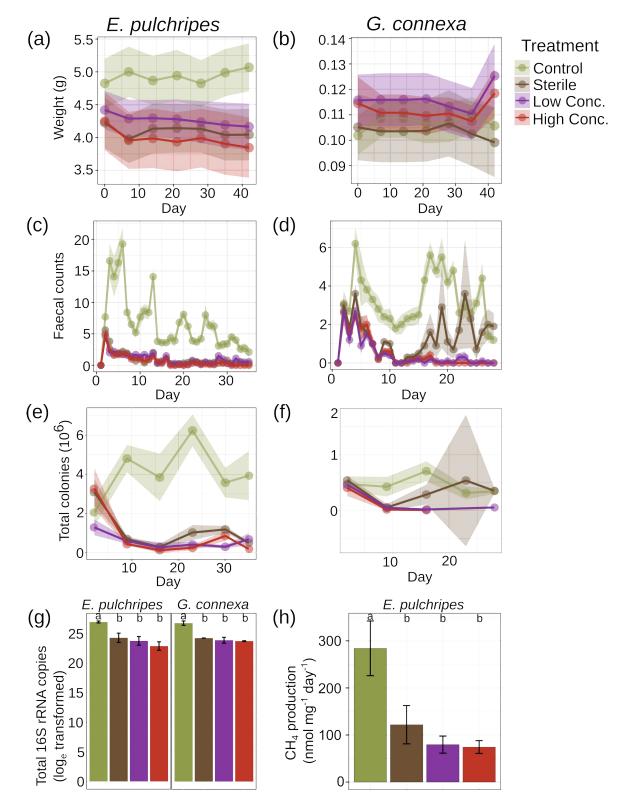
- 45. Gijzen HJ. 1991. Methanogenic bacteria as endosymbionts of the ciliate *Nyctotherus ovalis* in the Cockroach Hindgut. Appl Environ Microbiol 57:5.
- Zhou Z, Meng Q, Yu Z. 2011. Effects of methanogenic inhibitors on methane production and abundances of methanogens and cellulolytic bacteria in in vitro ruminal cultures. Appl Environ Microbiol 77:2634–2639.
- 47. Pereira AM, de Lurdes Nunes Enes Dapkevicius M, Borba AES. 2022. Alternative pathways for hydrogen sink originated from the ruminal fermentation of carbohydrates: Which microorganisms are involved in lowering methane emission? Anim Microbiome 4:5.
- 48. Husseneder C. 2010. Symbiosis in Subterranean Termites: A Review of Insights From Molecular Studies. Environmental Entomology 39:378–388.
- 49. Messer AC, Lee MJ. 1989. Effect of chemical treatments on methane emission by the hindgut microbiota in the termite *Zootermopsis angusticollis*. Microb Ecol 18:275–284.
- 50. Schneider T, Keiblinger KM, Schmid E, Sterflinger-Gleixner K, Ellersdorfer G, Roschitzki B, Richter A, Eberl L, Zechmeister-Boltenstern S, Riedel K. 2012. Who is who in litter decomposition? Metaproteomics reveals major microbial players and their biogeochemical functions. ISME J 6:1749–1762.
- 51. Tláskal V, Voříšková J, Baldrian P. 2016. Bacterial succession on decomposing leaf litter exhibits a specific occurrence pattern of cellulolytic taxa and potential decomposers of fungal mycelia. FEMS Microbiol Ecol 92:fiw177.
- 52. Purahong W, Wubet T, Lentendu G, Schloter M, Pecyna MJ, Kapturska D, Hofrichter M, Krüger D, Buscot F. 2016. Life in leaf litter: novel insights into community dynamics of bacteria and fungi during litter decomposition. Mol Ecol 25:4059–4074.
- 53. Voříšková J, Baldrian P. 2013. Fungal community on decomposing leaf litter undergoes rapid successional changes. 3. ISME J 7:477–486.
- 54. Geib SM, Filley TR, Hatcher PG, Hoover K, Carlson JE, Jimenez-Gasco M del M, Nakagawa-Izumi A, Sleighter RL, Tien M. 2008. Lignin degradation in wood-feeding insects. Proc Natl Acad Sci 105:12932–12937.
- 55. Nunez FS, Crawford CS. 1976. Digestive enzymes of the desert millipede Orthoporus ornatus (Girard) (Diplopoda: Spirostreptidae). Comp Biochem Physiol Part A Mol Integr Physiol 55:141–145.
- 56. Hackstein JH, Stumm CK. 1994. Methane production in terrestrial arthropods. Proc Natl Acad Sci 91:5441–5445.

- 57. Šustr V, Šimek M, Faktorová L, Macková J, Tajovský K. 2020. Release of greenhouse gases from millipedes as related to food, body size, and other factors. Soil Biol Biochem 144:107765.
- 58. Alonso-Pernas P, Bartram S, Arias-Cordero EM, Novoselov AL, Halty-deLeon L, Shao Y, Boland W. 2017. In Vivo Isotopic Labeling of Symbiotic Bacteria Involved in Cellulose Degradation and Nitrogen Recycling within the Gut of the Forest Cockchafer (*Melolontha hippocastani*). Front Microbiol 8:1970.
- 59. Sariyildiz T, Anderson JM. 2005. Variation in the chemical composition of green leaves and leaf litters from three deciduous tree species growing on different soil types. Forest Ecology and Management 210:303–319.
- 60. Horváthová T, Babik W, Bauchinger U. 2016. Biofilm feeding: microbial colonization of food promotes the growth of a detritivorous arthropod. ZooKeys 25.
- 61. Ihnen K, Zimmer M. 2008. Selective consumption and digestion of litter microbes by *Porcellio scaber* (Isopoda: Oniscidea). Pedobiologia 51:335–342.
- 62. Maraun M, Scheu S. 1996. Changes in microbial biomass, respiration and nutrient status of beech (*Fagus sylvatica*) leaf litter processed by millipedes (*Glomeris marginata*). Oecologia 107:131–140.
- 63. Semenyuk II, Tiunov AV. 2019. Foraging behaviour as a mechanism for trophic niche separation in a millipede community of southern Vietnam. Eur J Soil Biol 90:36–43.
- 64. Byzov BA, Thanh VN, Bab'Eva IP, Tretyakova EB, Dyvak IA, Rabinovich YM. 1998.
  Killing and hydrolytic activities of the gut fluid of the millipede *Pachyiulus flavipes* C.L. koch on yeast cells. Soil Biol Biochem 30:1137–1145.
- Joly F-X, Coq S, Coulis M, David J-F, Hättenschwiler S, Mueller CW, Prater I, Subke J-A.
  2020. Detritivore conversion of litter into faeces accelerates organic matter turnover.
  Communications Biology 3:660.
- Moritz L, Borisova E, Hammel JU, Blanke A, Wesener T. 2022. A previously unknown feeding mode in millipedes and the convergence of fluid feeding across arthropods. Sci Adv 8:eabm0577.
- 67. Hammer TJ, Bowers MD. 2015. Gut microbes may facilitate insect herbivory of chemically defended plants. Oecologia 179:1–14.
- 68. So WL, Nong W, Xie Y, Baril T, Ma H, Qu Z, Haimovitz J, Swale T, Gaitan-Espitia JD, Lau KF. 2022. Myriapod genomes reveal ancestral horizontal gene transfer and hormonal gene loss in millipedes. Nat Commun 13:3010.
- 69. Gerstaecker A. 1873. Die gliederthier-fauna des Sansibar-gebietes. CF Winter.

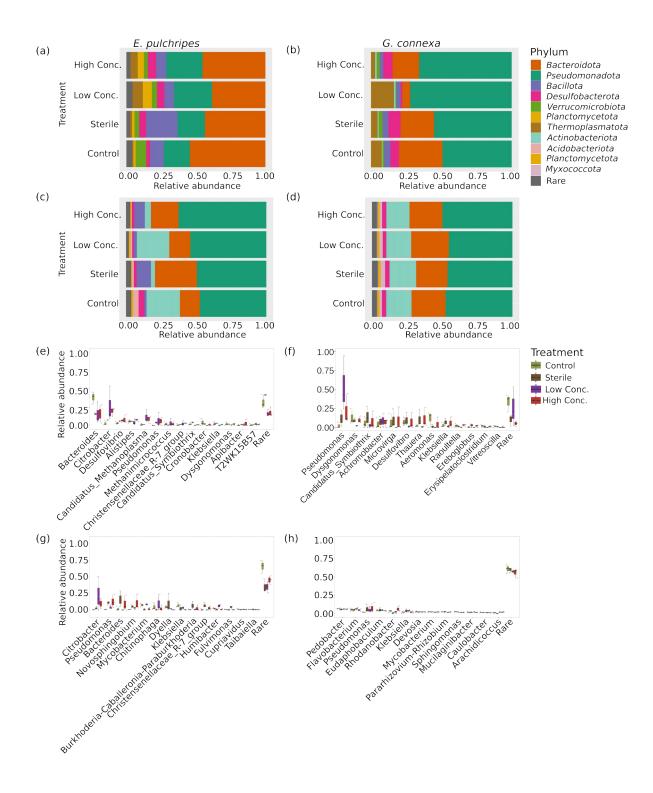
- 70. Kocourek P, Tajovský K, Dolejš P. 2017. New species of millipedes occurring in the Czech Republic: species discovered in the period 2003–2017 5.
- Zimmer M, Bartholmé S. 2003. Bacterial endosymbionts in *Asellus aquaticus* (Isopoda) and *Gammarus pulex* (Amphipoda) and their contribution to digestion. Limnol Oceanogr 48:2208–2213.
- Hayes JM. 2004. An introduction to isotopic calculations. Woods Hole Oceanographic Institution, Woods Hole, MA. https://www.whoi.edu/cms/files/jhayes/2005/9/IsoCalcs30Sept04\_5183.pdf. Retrieved 11 October 2023.
- 73. Angel R, Petrova E, Lara-Rodriguez A. 2021. Total nucleic acids extraction from soil V.6. protocols.io 6.
- 74. Angel R, Petrova E, Lara A. 2020. RNA-Stable Isotope Probing V.3. protocols.io 3.
- 75. Yu Y, Lee C, Kim J, Hwang S. 2005. Group-specific primer and probe sets to detect methanogenic communities using quantitative real-time polymerase chain reaction. Biotechnol Bioeng 89:670–679.
- 76. Angel R, Matthies D, Conrad R. 2011. Activation of methanogenesis in arid biological soil crusts despite the presence of oxygen. "PLoS One 6:e20453.
- 77. Liu CM, Kachur S, Dwan MG, Abraham AG, Aziz M, Hsueh P-R, Huang Y-T, Busch JD, Lamit LJ, Gehring CA. 2012. FungiQuant: a broad-coverage fungal quantitative real-time PCR assay. BMC Microbiol 12:1–11.
- 78. Naqib A, Poggi S, Green SJ. 2019. Deconstructing the polymerase chain reaction II: an improved workflow and effects on artifact formation and primer degeneracy. PeerJ 7:e7121.
- 79. Walters W, Hyde ER, Berg-Lyons D, Ackermann G, Humphrey G, Parada A, Gilbert JA, Jansson JK, Caporaso JG, Fuhrman JA, Apprill A, Knight R. 2016. Improved bacterial 16S rRNA gene (V4 and V4-5) and fungal internal transcribed spacer marker gene primers for microbial community surveys. mSystems 1:e00009-15.
- 80. RCore T. 2016. R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria.
- Bates D, M\u00e4chler M, Bolker B, Walker S. 2015. Fitting linear mixed-effects models using lme4. J Stat Softw 67:1–48.
- 82. Girden ER. 1992. ANOVA: Repeated measures. sage.
- Keselman HJ, Rogan JC. 1977. The Tukey multiple comparison test: 1953–1976. Psychol Bull 84:1050.

- 84. Goel MK, Khanna P, Kishore J. 2010. Understanding survival analysis: Kaplan-Meier estimate. Int J Ayurveda Res 1:274–278.
- 85. Martin M. 2011. Cutadapt removes adapter sequences from high-throughput sequencing reads. 1. EMBnet.journal 17:10–12.
- Callahan BJ, Sankaran K, Fukuyama JA, McMurdie PJ, Holmes SP. 2016. Bioconductor Workflow for Microbiome Data Analysis: from raw reads to community analyses. F1000Res 5:1492.
- 87. Quast C, Pruesse E, Yilmaz P, Gerken J, Schweer T, Yarza P, Peplies J, Glöckner FO. 2013. The SILVA ribosomal RNA gene database project: improved data processing and web-based tools. Nucleic Acids Res 41:D590-596.
- Davis NM, Proctor DM, Holmes SP, Relman DA, Callahan BJ. 2018. Simple statistical identification and removal of contaminant sequences in marker-gene and metagenomics data. Microbiome 6:226.
- 89. McMurdie PJ, Holmes S. 2013. phyloseq: An R package for reproducible interactive analysis and graphics of microbiome census data. PLOS One 8:e61217.
- Dixon P. 2003. VEGAN, a package of R functions for community ecology. J Veg Sci 14:927–930.
- 91. McKight PE, Najab J. 2010. Kruskal-wallis test. The corsini encyclopedia of psychology 1–
  1.
- 92. Dinno A, Dinno MA. 2017. Package 'dunn. test.' CRAN Repos 10:1–7.
- 93. Benjamini Y, Hochberg Y. 1995. Controlling the false discovery rate: a practical and powerful approach to multiple testing. J R Stat Soc, Ser B, Methodol 57:289–300.
- 94. Mangiafico S, Mangiafico MS. 2017. Package 'rcompanion.' Cran Repos 20:1–71.
- 95. Anderson MJ, Willis TJ. 2003. Canonical analysis of principal coordinates: a useful method of constrained ordination for ecology. Ecol 84:511–525.
- 96. Anderson MJ. 2001. A new method for non-parametric multivariate analysis of variance. Austral Ecol 26:32–46.
- 97. Martinez Arbizu P. 2020. pairwiseAdonis: Pairwise multilevel comparison using adonis. R package version 04 1.
- Lin H, Peddada SD. 2020. Analysis of compositions of microbiomes with bias correction. Nat Commun 11:3514.
- Angel R. 2019. Stable isotope probing techniques and methodological considerations using
   <sup>15</sup>N, p. 175–187. *In* Methods in Molecular Biology: Stable Isotope Probing. Springer.

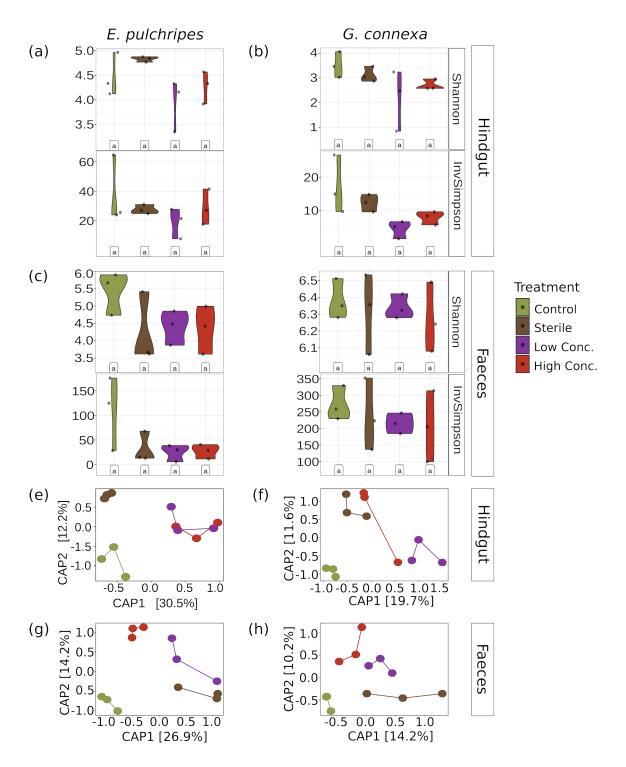
- 100. Pruesse E, Peplies J, Glöckner FO. 2012. SINA: accurate high-throughput multiple sequence alignment of ribosomal RNA genes. Bioinformatics 28:1823–1829.
- 101. Minh BQ, Schmidt HA, Chernomor O, Schrempf D, Woodhams MD, Von Haeseler A, Lanfear R. 2020. IQ-TREE 2: new models and efficient methods for phylogenetic inference in the genomic era. Mol Biol Evol 37:1530–1534.
- 102. Love MI, Huber W, Anders S. 2014. Moderated estimation of fold change and dispersion for RNA-seq data with DESeq2. Genome Biol 15:550.
- 103. Zhu A, Ibrahim JG, Love MI. 2019. Heavy-tailed prior distributions for sequence count data: removing the noise and preserving large differences. Bioinform 35:2084–2092.



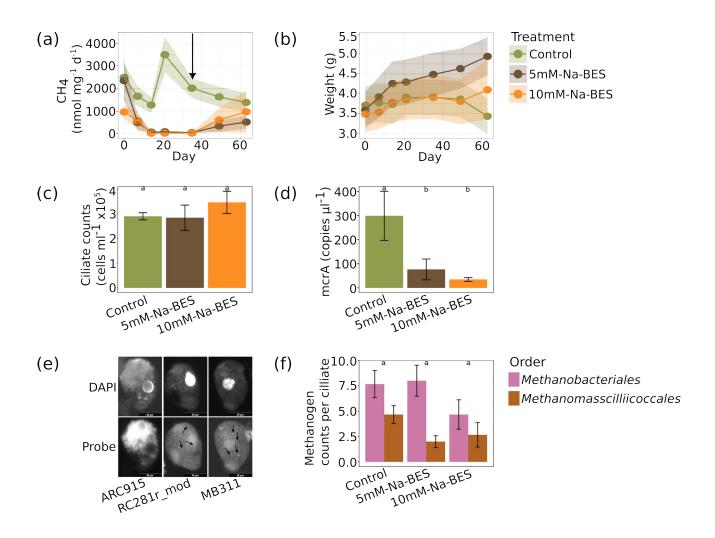
**Fig. 1. Effect of antibiotic treatment on** *E. pulchripes* and *G. connexa*. Time series of mean weight loss (mean  $\pm$  SE ribbon) in (a) *E. pulchripes* and (b) *G. connexa*; faecal counts in (c) *E. pulchripes* and (d) *G. connexa*; total colony forming units in (e) *E. pulchripes* and (f) *G. connexa*; (g) 16S rRNA gene copy numbers in the faeces; and (h) CH<sub>4</sub> production rate after 35 days of antibiotics treatment in *E. pulchripes*. 'High Conc.' and 'Low Conc.' refer to the concentration of applied antibiotics (see Materials and Methods for more details). Different lower case letters in panels g and h denote statistical significance. See Results for a detailed description of the statistical tests performed on the time series (panels a-f).



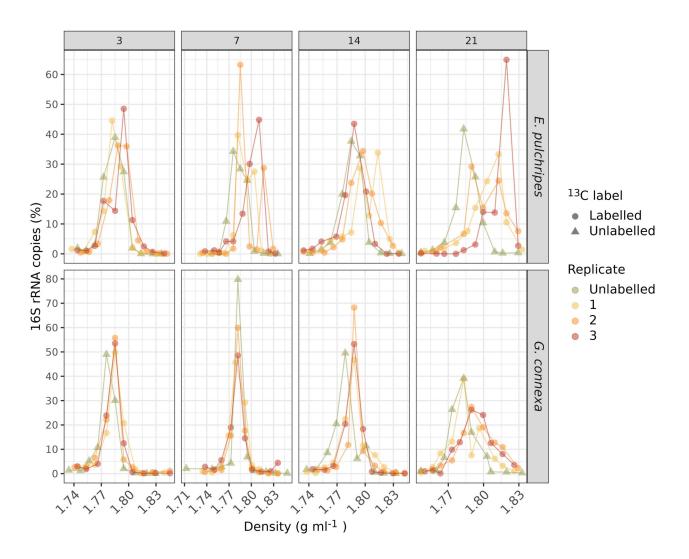
**Fig. 2. Effect of antibiotic treatment on the taxonomic composition of prokaryotes in** *E. Pulchripes* (left) and *G. Connexa* (right) following treatment. Phylum distribution in the hindguts (a and b) and the faeces (c and d). Distribution at genus level in the hindguts (e and f) and faeces (g and h). 'High Conc.' and 'Low Conc.' refer to the conc. of antibiotics applied (see Materials and Methods for more details).



**Fig. 3.** Effect of antibiotic treatment on the alpha and beta diversity indices of the microbial communities in the hindgut and faeces in *E. pulchripes* (left) and *G. connexa* (right). Alpha diversity values for each species, stratified by treatment groups for hindgut (a and b) and faeces samples (c and d) from *E. pulchripes* and *G. connexa*. The statistical test was based on Kruskal–Wallis (identical letters denote p >0.05). Dissimilarity between hindgut (e and f) and faeces (g and h) microbial communities in the different treatments using constrained principal coordinates analysis (PcoA) with the model Dist.Mat ~ Treatment for each species and sample type separately.

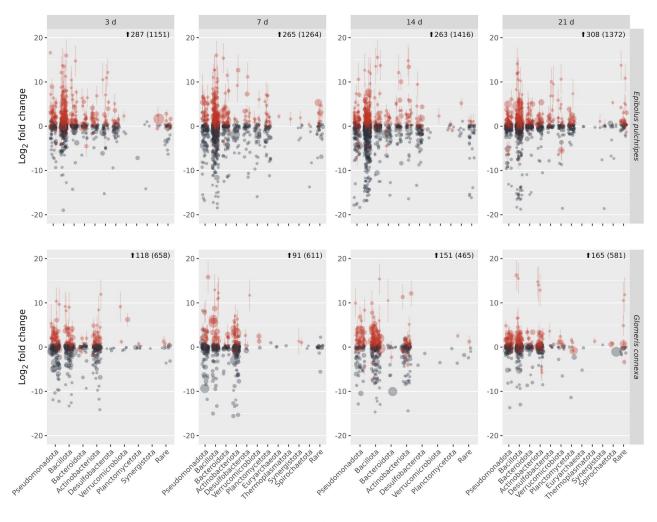


**Fig. 4. Effect of BES treatment on CH**<sub>4</sub> **emissions from** *E. pulchripes*, **animal weight**, **ciliates and ciliate-associated methanogens.** (a) Emission rates of CH<sub>4</sub> over time followed by recommence of methane production after the switch to untreated litters (indicated by the arrow). (b) Change in the weight of *E. pulchripes* over time. (c) Enumeration of symbiotic ciliates found in the faeces following BES treatment. (d) *mcrA* gene copy numbers in the faecal samples following BES treatment. (e) Fluorescence microscopy images of ciliates and the two most-abundant endosymbiotic methanogens in faecal samples of *E. pulchripes* using DAPI and CARD-FISH probes. ARC915: general archaea, RC281r\_mod: *Methanomassciillicoccales*, and MB311: *Methanobacteriales* in the 10mM-Na-BES-treated group. (f) Enumeration of the methanogens associated with ciliates using FISH signals.



**Fig. 5. Bacterial 16S rRNA copies recovered from each fraction in the SIP gradients.** rRNA copies relative to the total number of rRNA copies obtained from the entire gradient against the buoyant density of each fraction. Labelled RNA is expected to be found in fractions with density >1.795 g ml<sup>-1</sup>.

bioRxiv preprint doi: https://doi.org/10.1101/2024.03.01.582937; this version posted March 4, 2024. The copyright holder for this preprint (which was not certified by peer review) is the author/funder, who has granted bioRxiv a license to display the preprint in perpetuity. It is made available under aCC-BY-NC 4.0 International license.



Mean abundance (%) 1.0 1.9 2.9 3.8 Significance at Fail Pass p < 0.1

Fig. 6. Differentially abundant ASVs between the labelled and unlabelled gradients of the SIP experiments. Comparison of the relative abundance of each ASV from *E. pulchripes* and *G. connexa*. Each subfigure represents a triplicate. The plot shows the most abundant phyla in the dataset in decreasing abundance. The differential abundance of any particular ASV is given in  $Log_2$  fold change. "Rare" indicates phyla with mean relative abundance below 0.45%.